

# **JGR** Biogeosciences

# **RESEARCH ARTICLE**

10.1029/2019JG005613

#### **Key Points:**

- Model reveals spatial variability in seasonality of nitrous oxide emissions driven by differences in stratification intensity and duration
- The northwest European shelf contributes between 3.3% and 6.8% to total nitrous oxide emissions from European shelves and estuaries
- Total amount of nitrous oxide is controlled by ammonium availability; changes in nutrient input can alter regional emissions

#### **Supporting Information:**

- Supporting Information S1
- Figure S1
- Figure S2

#### Correspondence to:

G. Lessin, gle@pml.ac.uk

#### Citation:

Lessin, G., Polimene, L., Artioli, Y., Butenschön, M., Clark, D. R., Brown, I., & Rees, A. P. (2020). Modeling the seasonality and controls of nitrous oxide emissions on the northwest European continental shelf. *Journal of Geophysical Research: Biogeosciences*, *125*, e2019JG005613. https://doi.org/ 10.1029/2019JG005613

Received 16 DEC 2019 Accepted 2 MAY 2020 Accepted article online 8 MAY 2020

#### **Author Contributions:**

**Conceptualization:** Gennadi Lessin, Luca Polimene, Yuri Artioli, Momme Butenschön **Data curation:** Ian Brown, Andrew P.

Rees Formal analysis: Gennadi Lessin, Luca Polimene, Yuri Artioli, Momme Butenschön, Darren R. Clark, Ian Brown, Andrew P. Rees Investigation: Gennadi Lessin, Luca Polimene, Yuri Artioli, Momme (continued)

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# Modeling the Seasonality and Controls of Nitrous Oxide Emissions on the Northwest European Continental Shelf

Gennadi Lessin<sup>1</sup>, Luca Polimene<sup>1</sup>, Yuri Artioli<sup>1</sup>, Momme Butenschön<sup>2</sup>, Darren R. Clark<sup>1</sup>, Ian Brown<sup>1</sup>, and Andrew P. Rees<sup>1</sup>

<sup>1</sup>Plymouth Marine Laboratory, Plymouth, UK, <sup>2</sup>Euro-Mediterranean Center on Climate Change (CMCC), Bologna, Italy

**Abstract** Estimates of oceanic emissions of nitrous oxide ( $N_2O$ ) are surrounded by a considerable degree of uncertainty, particularly regarding the contribution of productive shelf regions, where assessments are based on limited observations. In this paper, we have applied a coupled hydrodynamic-biogeochemical model resolving  $N_2O$  dynamics to estimate  $N_2O$  emissions within the northwest European continental shelf. Based on 10-year average distributions (2006–2015), dominant seasonal patterns of  $N_2O$  air-sea exchange were identified. Within the southwest region of the shelf and deep parts of the North Sea, emissions are highest during winter. Peak emissions during late autumn are typical for the northwest part of the shelf and central North Sea, while in the western English Channel, Irish Sea and western North Sea peak outflux shifts toward early autumn. Within these regions, most  $N_2O$  production occurs below the seasonal pycnocline, and duration and intensity of stratification defines the timing and rate of its subsequent release to the atmosphere. In contrast, within the southeast North Sea and most of the coastal areas, lack of stratification allows the excess  $N_2O$  to outgas as soon as it is produced, driven by ammonium availability, resulting in peak emissions in summer. We estimate that  $N_2O$  emissions from the northwest European shelf contribute 0.02224 Tg N to the atmosphere annually, that is, between 3.3–6.8% of total emissions from European shelves and estuaries.

# 1. Introduction

Nitrous oxide ( $N_2O$ ) is the third most important long-lived greenhouse gas contributing ~6% of the direct anthropogenic radiative forcing, preceded by carbon dioxide and methane which contribute ~76% and ~16%, respectively (IPCC, 2013). With a global warming potential about 300 times higher than that of carbon dioxide,  $N_2O$  plays a significant role in the greenhouse effect (Myhre et al., 2013).  $N_2O$  also contributes to the destruction of stratospheric ozone (Andreae & Crutzen, 1997; Nevison & Holland, 1997), its emissions currently having the largest ozone depletion potential of all ozone depleting substances (Carpenter et al., 2014; Ravishankara et al., 2009).

The main pathways for natural  $N_2O$  emissions to the atmosphere are production during microbially mediated nitrification and denitrification processes within soils and in the ocean. During nitrification,  $N_2O$  is produced as a by-product of ammonium oxidation to hydroxylamine, or via nitrite reduction, with enhanced yield at oxygen-limited conditions (Capone, 1991; Goreau et al., 1980; Löscher et al., 2012).  $N_2O$  is also an intermediate product of denitrification, leading to either net production or net consumption flux depending on oxygen conditions (Bianchi et al., 2012; Cohen & Gordon, 1978; Nevison et al., 2003).

The ocean is an important source of  $N_2O$  to the atmosphere. According to the Fifth Assessment Report of Intergovernmental Panel on Climate Change, in 2006 mean total emissions of  $N_2O$  to the atmosphere from all natural and anthropogenic sources combined were 17.9 Tg N yr<sup>-1</sup>, of which oceans contributed 21.2% (3.8 Tg N yr<sup>-1</sup>), while 3.4% (0.6 Tg N yr<sup>-1</sup>) originated from coastal areas including estuaries and rivers. However, these estimates are surrounded by a considerable degree of uncertainty with ranges of global, oceanic, and coastal emissions between 8.1–30.7, 1.8–9.4<sup>o</sup> and 0.1–2.9 Tg N yr<sup>-1</sup>, respectively (Ciais et al., 2013). Bange et al. (1996) suggested that coastal and upwelling regions which are known for high N<sub>2</sub>O production were

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Butenschön, Darren R. Clark, Ian Brown, Andrew P. Rees

Methodology: Gennadi Lessin, Luca Polimene, Yuri Artioli, Momme Butenschön, Darren R. Clark, Ian Brown, Andrew P. Rees Resources: Ian Brown, Andrew P. Rees Software: Gennadi Lessin, Luca Polimene, Yuri Artioli, Momme Butenschön Validation: Gennadi Lessin, Ian Brown Andrew P. Rees

Visualization: Gennadi Lessin Writing - original draft: Gennadi Lessin, Luca Polimene, Yuri Artioli, Momme Butenschön, Darren R. Clark, Ian Brown, Andrew P. Rees

Writing – review & editing: Gennadi Lessin, Luca Polimene, Yuri Artioli, Momme Butenschön, Darren R. Clark, Ian Brown, Andrew P. Rees poorly represented in previous budgets and provided a revised estimate for global oceanic production rate of  $4.5-7.64 \text{ Tg N yr}^{-1}$ .

Our current ability for more accurate estimates of  $N_2O$  distributions and emissions is hindered by the lack of comprehensive, quality-controlled observations able to describe the spatial and temporal variability, especially within dynamic marginal seas and coastal areas. International intercomparison exercises, such as the one conducted by Wilson et al. (2018) will assure consistent quality control of  $N_2O$  measurements, and together with other recent international initiatives, such as the MarinE MethanE and NiTrous Oxide database (MEMENTO) (Bange et al., 2009; Kock & Bange, 2015) and the proposed  $N_2O$  Observation Network for harmonized measurements from various platforms (Bange et al., 2019) will contribute toward improved quality and spatiotemporal coverage of measurement data, essential for the reduction of uncertainties in global and regional estimates of emissions and their variability, as well as further development and validation of numerical models of  $N_2O$ .

Coupled hydrodynamic-biogeochemical models integrate our current understanding of ocean functioning, enabling the analysis, interpretation, and upscaling of available observational data, as well as predicting the response of N<sub>2</sub>O to natural and anthropogenic stressors and changing climate. Recent modeling efforts largely focused on assessing mechanisms of N2O production and projection of emissions at a global scale (Ji et al., 2018; Martinez-Rey et al., 2015; Suntharalingam et al., 2000; Suntharalingam & Sarmiento, 2000; Zamora & Oschlies, 2014). However, global models currently poorly represent coastal and shelf seas despite their disproportionally large role in biogeochemical cycling (Holt et al., 2017). To reduce the current uncertainties related to N<sub>2</sub>O dynamics within coastal regions and their contribution to global emission rates, regional models with higher resolution of local hydrography and a detailed description of biogeochemistry are needed. These models can provide information highly relevant for the support of regional marine environmental policy development and management (Hyder et al., 2015). Furthermore, regional models allow the analysis of drivers controlling spatiotemporal distributions and sea-to-air fluxes of N<sub>2</sub>O within shelf sea regions and therefore can help in the interpretation of existing data sets and inform the design of observational efforts to further improve understanding of N2O processes and its emission estimates. To this end, we have applied a complex biogeochemical-ecological model resolving N<sub>2</sub>O dynamics on the northwest European continental shelf, with the aim to (a) identify dominant seasonal patterns of  $N_2O$  air-sea flux and the drivers underpinning its variability, and (b) quantify the "present-day" contribution of the northwest European shelf as a source of N<sub>2</sub>O to the atmosphere.

# 2. Materials and Methods

# 2.1. Model Description and Setup

For the purpose of our study we use the 3-D coupled hydrodynamic-biogeochemical modeling system NEMO-FABM-ERSEM configured on the Atlantic Meridional Margin (AMM7) domain, which is based on an update of Edwards et al. (2012). The geographical domain extends from  $20^{\circ}$ W to  $13^{\circ}$ E and  $40^{\circ}$ N to  $65^{\circ}$ N (Figure 1), and has a horizontal resolution of  $1/15^{\circ}$  in latitudinal and  $1/9^{\circ}$  in longitudinal direction, corresponding to  $\sim$ 7 km. Vertically, the model resolution was improved to  $50 \sigma$ -z layers, from the original 32, and has a minimum bathymetry of 10 m. In areas of steep bathymetry change the total number of layers is decreased to reduce spurious vertical transport.

Atmospheric forcing is provided by the ERA-interim reanalysis of the European Centre for Medium-Range Weather Forecasts (Dee et al., 2011), while physical oceanic boundary conditions are taken from GLORYS reanalysis (Ferry et al., 2012). The model is initialized on 1 January 2000 with the physical and biogeochemical states from a previous long-term hindcast simulation and a homogeneous concentration of  $N_2O$  of 10.75 nmol  $L^{-1}$ , based on typical observational values for the study area. This combination of initial conditions allowed for a short 6-year spin-up period (2000–2005), with the last 10 years (2006–2015) used for the assessment of  $N_2O$  dynamics.

Formulation and parameterization of the biogeochemical-ecological model ERSEM follows that of Butenschön et al. (2016), extended with the processes of  $N_2O$  production and air-sea exchange described below, implemented within FABM—Framework for Aquatic Biogeochemical Models (Bruggeman & Bolding, 2014).





**Figure 1.** Map of the model domain. Black contour line outlines the northwest European continental shelf area (part of the modeled domain delineated with 200 m isobath, as well as Norwegian trench). Shaded area defines the extent of the on -shelf part of the UK Exclusive Economic Zone (EEZ). Red box delimits the mapped area shown on Figure 2, where measurements of  $N_2O$  were taken within the UK Shelf Sea biogeochemistry program (cruises DY018, DY029, and DY033).

In oxygenated waters N<sub>2</sub>O is mainly produced as a by-product during the process of nitrification (Freing et al., 2012). Hence, the pelagic production of N<sub>2</sub>O ( $N_2O_{prod}$ , nmol N L<sup>-1</sup> day<sup>-1</sup>) was parameterized as proportional to total nitrification flux, assuming enhancement under low oxygen conditions:

$$N_2 O_{\text{prod}} = \alpha F_{\text{nit}} Enh, \tag{1}$$

where  $\alpha$  is the N<sub>2</sub>O yield constant,  $F_{\text{nit}}$  is the total nitrification flux (µmol N L<sup>-1</sup> day<sup>-1</sup>), and *Enh* is a factor enhancing N<sub>2</sub>O production at low-oxygen conditions, parameterized as

$$Enh=\min\left(E_{\max},\frac{Ox^3+h_{ox}}{Ox^3}\right),\tag{2}$$

where  $E_{\text{max}}$  is the maximum value of enhancement, Ox is the oxygen concentration (µmol O<sub>2</sub> L<sup>-1</sup>) and  $h_{ox}$  is a constant. The value of  $E_{\text{max}}$  of 20 was chosen following Codispoti (2010), while  $h_{ox}$  of 2,700 (µmol L<sup>-3</sup>)<sup>3</sup> was applied.

Nitrification flux  $F_{nit}$  is modeled implicitly as a first-order process dependent on availability of ammonium (*Amm*) and modified by temperature, oxygen and pH:

$$F_{\rm nit} = r_{\rm nit} f(T) f(Ox) f(pH) Amm$$
(3)

Maximum ammonium mass-specific nitrification rate  $r_{nit}$  of 0.5 day<sup>-1</sup> was applied. The value is rather



close to the mean specific nitrification rate of  $0.55 \text{ day}^{-1}$  reported by Yool et al. (2007) based on a wide range of measurements dominated by data from the northeast region of Atlantic Ocean.

Temperature dependency of nitrification is modeled following Blackford et al. (2004):

$$f(T) = Q_{10}^{\frac{T-10}{10}} - Q_{10}^{\frac{T-32}{3}}$$
(4)

where T is water temperature in degrees Celsius and  $Q_{10}$  is a coefficient of 2.0.

Dependency of nitrification on oxygen is parameterized as

$$f(Ox) = \frac{Ox^3}{Ox^3 + h_{ox}}$$
(5)

Finally, limitation of nitrification by pH is accounted for as

$$f(pH) = \min(2, \max(0, 0.6111pH - 3.8889))$$
(6)

The formulation was parameterized by Blackford and Gilbert (2007) based on observational work of Huesemann et al. (2002).

According to Bange (2008), reported  $N_2O$  yields from nitrification range from 0.004% to 0.4%. Dore and Karl (1996) reported a higher yield of 0.5% for the North Pacific time series station ALOHA, while Ji et al. (2018) derived yield uncertainty range of 0.04–0.12% at oxygen concentrations above 50  $\mu$ mol L<sup>-1</sup>, based on measurements in the Eastern Tropical Pacific. In our model we implemented a value of 0.1% for the N<sub>2</sub>O production yield at nitrification, which allowed realistic distributions of N<sub>2</sub>O to be modeled on the northwest European Shelf.

Several modeling studies exclude  $N_2O$  production in the photic zone (e.g., Battaglia & Joos, 2018; Martinez-Rey et al., 2015) following an assumption that nitrification by ammonia-oxidizing bacteria (AOB) is inhibited by light (Horrigan & Springer, 1990; Ward et al., 1982). However, there is a growing evidence that nitrification by ammonia-oxidizing archaea (AOA) takes place within the photic zone, and that this mechanism might be the dominant route for ammonia oxidation in the upper ocean (e.g., Beman et al., 2012; Newell et al., 2013; Peng et al., 2016; Ward, 2008), although photosensitivity of AOA has also been reported (Horak et al., 2018; Merbt et al., 2012). Several studies summarized by Yool et al. (2007) (see Figure 1b therein) have shown that there is no clear increase in nitrification rate with depth, which would indicate inhibition by light in the euphotic zone. Other studies have also demonstrated nitrification to take place within the photic zone (Clark et al., 2008; Dore & Karl, 1996; Wankel et al., 2007). Moreover, experiments conducted by Smith et al. (2014) demonstrated nitrification rates in the photic zone to be more strongly regulated by competition with phytoplankton for ammonium rather than directly by light. Since details on sensitivity of nitrification to light intensity are still debated, within this work we followed the original formulation of ERSEM, where nitrification rate is not affected by light.

Pelagic denitrification-related sources and sinks of  $N_2O$  were not included in the present model formulation. The majority of waters within the study area are well oxygenated, and for denitrification to take place an anoxic environment is required. Therefore, we assume that the amount of denitrification occurring in the system is negligible, as was previously confirmed by in situ observations in the North Sea (Law & Owens, 1990). Furthermore, we did not include benthic sources and sinks of  $N_2O$ in the model formulation applied here based on the evidence that  $N_2O$  production rate in the North Sea sediments is very low (Law & Owens, 1990). It is likely for the major pathway of nitrogen removal in the sediments within our study area to be not denitrification, but anammox, as found by Kitidis et al. (2017) for the sediments of the Celtic Sea. The process of anammox does not involve  $N_2O$  production or consumption.

Air-sea flux of nitrous oxide ( $F_{N_2O}$ , µmol N<sub>2</sub>O m<sup>-2</sup> day<sup>-1</sup>) was calculated as a product of the gas-transfer velocity  $k_{N_2O}$  and the difference between equilibrium and surface water gas concentrations:



$$F_{N_2O} = k_{N_2O} \left( N_2 O_{eq} - N_2 O_{sea} \right)$$
<sup>(7)</sup>

The gas transfer velocity was calculated following Wanninkhof (1992):

$$k_{\rm N_20} = 0.39 U^2 \left(\frac{Sc}{660}\right)^{-0.5},$$
 (8)

where *U* is the wind speed and *Sc* is the Schmidt number. Solubility of  $N_2O$  in seawater was calculated according to Weiss and Price (1980). Time-varying atmospheric partial pressure of  $N_2O$  based on monthly values over marine surface sites for the Northern Hemisphere (NOAA, 2014) was applied uniformly over the model domain.

#### 2.2. Self-Organizing Maps

To identify dominant patterns in seasonality of  $N_2O$  air-sea exchange on the northwest European shelf we employed self-organizing map (SOM). SOM is a type of artificial neural network that uses unsupervised learning to retrieve the main underlying patterns or structures (termed nodes) within a complex data set and to produce its low-dimensional discretized representation (Kohonen, 2001). The number of SOM nodes is somewhat subjective and depends on the characteristics of input data and research questions. Following experimentation with different numbers of SOM nodes, we chose to base our implementation on four nodes organized in a linear pattern, which allowed for determination of distinct seasonal patterns of  $N_2O$  fluxes within small number of classes. Weights of each node were initialized as a linear combination of the first two principal components of the data. Thereafter, SOM was trained using all input data vectors sequentially. At the end of the training process, each node represented one of four distinct types of annual cycles of  $N_2O$  air-sea flux within the northwest European shelf. The annual cycles at each model grid point were then mapped to their closest node to define four regions with distinct seasonal pattern of  $N_2O$  fluxes. We use MiniSOM, a minimalistic NumPy based implementation of SOM (Vettigli, 2019).

#### 2.3. Observational Data

Measurements of  $N_2O$  were made during four research cruises in the study area. These were a circumnavigation of the UK during June and July 2011 on board *RRS Discovery* (UK Ocean Acidification program cruise D366) and three cruises on board *RRS Discovery* to the Celtic Sea in November 2014, April 2015, and July 2015 (UK Shelf Sea Biogeochemistry program cruises DY018, DY029, and DY033, respectively).

A total of 181, 62, 241, and 252 samples were collected during cruises D366, DY018, DY029, and DY033, respectively. At each station, up to 10 samples were collected and analyzed at depths between the sea surface and within 10 m of the seabed. Seawaters were collected into 1 L glass bottles using acid cleaned Tygon tubing directly from CTD Niskin bottles. Samples were overfilled in order to expel air bubbles, poisoned with 200  $\mu$ l of saturated HgCl<sub>2</sub> solution and temperature equilibrated at 25.0  $\pm$  0.5 °C. In all cases samples were analyzed within 8 hr of collection. N<sub>2</sub>O was determined by single-phase equilibration gas chromatography with electron capture detection similar to that described by Upstill-Goddard et al. (1996). Each individual sample was calibrated against three certified ( $\pm$ 5%) reference standards of 287, 402, and 511 ppb (Air Products Ltd) which are traceable to NOAA WMO-N<sub>2</sub>O-X2006A scale for N<sub>2</sub>O mole fractions. Mean instrument precision from daily, triplicate analyses of the three calibration standards was <0.95% coefficient of variation. Concentrations of N<sub>2</sub>O in seawater were calculated from solubility tables of Weiss and Price (1980) at equilibration temperature (~25°C) and salinity.

# 3. Results

#### 3.1. Model Validation

The model skill in reproducing seasonal variability in vertical distributions of  $N_2O$  was assessed by comparing modeled saturations with measurement data collected in the Celtic Sea in November 2014 (Rees, 2019a), April 2015 (Rees, 2019b) and July 2015 (Rees & Stephens, 2019). Covering different seasons and extending





**Figure 2.** Comparison of modeled (solid lines) and measured (red dots) vertical profiles of N<sub>2</sub>O saturation (%) in the Celtic Sea. Limits of standard deviation for modeled data are shown (dashed lines). For measurements, standard deviations are indicated if several measurements were taken at the same depth (horizontal lines). Maps (area denoted by red box on Figure 1) specify locations of stations sampled within each cruise.

geographically over a transect in the order of 600 km (Figure 2), these research expeditions captured contrasting conditions of the physical status and biological productivity of the shelf sea (e.g., García-Martín et al., 2019; Kitidis et al., 2017; Ruiz-Castillo et al., 2019).



Figure 3. The difference between modeled and measured N<sub>2</sub>O saturations (%) for June–July 2011, comparing measured values closest to the surface (<10 m depth) with corresponding results of the model. Positive (negative) values mean overestimation (underestimation) of N<sub>2</sub>O saturations by the model.

Vertical profiles of N<sub>2</sub>O saturations averaged over time and horizontal space were constructed for data collected within each of the three cruises and were compared with vertical profiles of modeled N<sub>2</sub>O for the corresponding locations and time instances. During November 2014 (Figure 2, left), the model showed very low variability and vertically uniform saturations of N2O close to equilibrium (about 101.2%) in the upper 50 m of the water column, and a pronounced increase in saturations between the depths of 50 and 70 m, up to about 120.2%. This increase was also evident in direct observations, though the model tends to slightly overestimate N2O saturation at depths greater than 60 m. Vertically uniform N2O saturations throughout water column were characteristic for the April 2015 (Figure 2, center), as shown both in the observational data and reproduced in the model, with the latter close to equilibrium, while the former mostly slightly undersaturated, with average saturations of 95.98% and 100.5% in the observations and in the model, respectively. For July 2015 (Figure 2, right), measurements showed N<sub>2</sub>O saturations mostly between 100% and 110%. The model showed mean saturation of 101.8% in the upper 10 m, which increased to a maximum of 113.9% at 40 m depth, followed by a steady decrease with greater depth, bringing modeled results back into agreement with the observed values.



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**Figure 4.** Seasonally averaged distribution of modeled  $N_2O$  air-sea fluxes ( $\mu$ mol m<sup>-2</sup> day<sup>-1</sup>) over the northwest European continental shelf based on a 2006–2015 climatology: Spring (March–May), summer (June–August), autumn (September–November), and winter (December–February). Positive values indicate outgassing of  $N_2O$  into the atmosphere, negative-uptake by the ocean. The red contour line separates regions with positive and negative fluxes.

The model skill in reproducing shelf-wide spatial distribution of N<sub>2</sub>O saturations within surface waters was assessed by comparing model results with measurement data collected between June and July 2011 (Brown & Rees, 2014) (also available from the MEMENTO database; Bange et al., 2009; Kock & Bange, 2015), covering a range of geographical locations with varying hydrographical and biogeochemical conditions within the northwest European shelf (Figure 3). Measurements taken closest to the surface (where they were taken at depth < 10 m) were compared with the corresponding values from the model. Modeled saturations generally match measurements rather well, although there is a slight tendency for overestimation compared to observational data (modeled and measured saturations averaged over all stations are 107.36  $\pm$  4.49% and 102.62  $\pm$  6.75%, respectively). The overestimation is especially apparent for the highly productive shallow region in the southern North Sea, as well as Kattegat area.

#### 3.2. Mean Seasonal Distributions of N<sub>2</sub>O Air-Sea Fluxes

Characteristic seasonality of air-sea fluxes of  $N_2O$  over the northwest European continental shelf area (defined here as part of the modeled domain delineated with the 200 m isobath, but also including the Norwegian Trench) is assessed based on a climatological year representing the average of a 10-year model simulation (2006–2015).

The northwest European continental shelf as a whole is a source of  $N_2O$  to the atmosphere, emitting on average 2.03  $\pm$  1.93 µmol  $N_2O$  m<sup>-2</sup> day<sup>-1</sup>. However, there is a considerable amount of variability in the distribution of fluxes between seasons, as well as spatial differences within each individual season (Figure 4).





**Figure 5.** (left) Seasonal patterns of N<sub>2</sub>O sea-to-air fluxes, mean over the region (black lines) and at selected stations (red lines). Positive values indicate outgassing to the atmosphere, negative-uptake. To remove weather-related noise, fluxes are shown as 14-day moving average values. (right) The corresponding spatial distribution of classes (A–D) within the northwest European continental shelf area. Location of stations CCS (Central Celtic Sea; 49.39°N,8.58°W), NNS (Northern North Sea; 57.25°W,2.5°E), LF (Long Forties area; 57.25°N,0.5°W), SNS (Southern North Sea; 53°N,2.5°E), and NT (Norwegian Trench; 58.75°W,4.5°E) are identified.

Fluxes to the atmosphere increase from spring toward summer, especially in the coastal areas around UK and the south-eastern North Sea region where maximum values rise from ~4  $\mu$ mol m<sup>-2</sup> day<sup>-1</sup> in spring to ~10  $\mu$ mol m<sup>-2</sup> day<sup>-1</sup> in summer. In autumn, in the North Sea, fluxes are comparatively higher offshore (reaching ~7  $\mu$ mol m<sup>-2</sup> day<sup>-1</sup>) than in coastal regions. Fluxes are lower than in summer in the southernmost part of the North Sea and eastern English Channel, but higher in the western English Channel and eastern part of the Celtic Sea. Also in the northern shelf regions, N<sub>2</sub>O outflux is greater in autumn than in summer. In winter, equilibrium conditions or weak negative fluxes (i.e., uptake of N<sub>2</sub>O from the atmosphere) prevail over the western part of the North Sea, the English Channel and the Irish Sea. The rest of the area is characterized by low outgassing of <3  $\mu$ mol m<sup>-2</sup> day<sup>-1</sup>, apart from a region west and southwest of Ireland (4–6  $\mu$ mol m<sup>-2</sup> day<sup>-1</sup>) and along the western coast of France (up to 9  $\mu$ mol m<sup>-2</sup> day<sup>-1</sup>).

# 3.3. Dominant Seasonal Patterns of N<sub>2</sub>O Air-Sea Fluxes

Through the application of SOM, dominant classes (A–D) of seasonal cycle in  $N_2O$  air-sea fluxes on the northwest European shelf were identified (Figure 5, left). Class A is characterized by high fluxes during winter months (November–December), while low fluxes are typical for the rest of the year. Within classes B and C peak fluxes shift toward late-autumn (October–November) and early-autumn (September–October) months, respectively. For all the above classes, a minor increase in fluxes from early spring toward May–June is characteristic and consecutively more pronounced. Class D is characterized by bell-shaped seasonal distribution with higher fluxes from late spring until early autumn, peaking in summer (July–August), and low fluxes throughout winter months.

Spatially, seasonal distributions corresponding to class A are found within the southwestern region of our study area (Celtic Sea, outer part of Armorican shelf) and offshore region of the northern North Sea (Figure 5, right). Class B fluxes are characteristic for central North Sea and outer parts of Irish and



Shetland Shelves. Temporal patterns corresponding to class C can be found in the western English Channel, Irish Sea, coastward region of Shetland Shelf and western part of North Sea adjacent to UK coast. Class D represents south-eastern North Sea, central English Channel, coastal areas around UK, Ireland, and France, as well as Norwegian Trench.

# 4. Discussion

# 4.1. Controls of Seasonality in N<sub>2</sub>O Air-Sea Fluxes

To determine the major drivers shaping the seasonal patterns of  $N_2O$  fluxes identified in the previous section, we have selected four stations within the model domain: CCS (Central Celtic Sea; 49.39°N,8.58°W), NNS (Northern North Sea; 57.25°W,2.5°E), LF (Long Forties area; 57.25°N,0.5°W), and SNS (Southern North Sea; 53°N,2.5°E), respectively representing classes A, B, C, and D (see Figure 5). Temporal patterns of air-sea fluxes of  $N_2O$ , as well as their absolute values at these stations rather well reflect the average-per-class values, showing short-lasting winter maximum for class A, bell-shaped seasonal distributions of class D (although with a maximum skewed toward late-spring period at SNS station), and classes B and C showing transitional dynamics between the former two (Figure 5, left).

The distinct seasonality of N<sub>2</sub>O air-sea fluxes among these stations, and, therefore, between the identified regions, depends on the seasonality of  $N_2O$  saturation state within surface waters and the underlying mechanisms controlling it. Thus, the region corresponding to class A is characterized by relatively deep water and strong seasonal density stratification (Figure 6; Station CCS). Excess N<sub>2</sub>O (i.e., saturation state >100%) builds up from late spring until late autumn below the seasonal pycnocline. Intrusion of these water masses, increasingly oversaturated with N<sub>2</sub>O, into the upper mixed layer, is effectively prevented until the late autumn when stratification collapses and excess N<sub>2</sub>O is allowed to escape into the atmosphere. A similar mechanism for the buildup of N<sub>2</sub>O oversaturation exists within class B (Figure 6; Station NNS), which is also characterized by strong seasonal density stratification, albeit at shallower water depths. During the stratified period, oversaturation of  $N_2O$  builds up below, but also within the pycnocline. As the water column gradually starts to mix from September onward, this facilitates earlier outgassing of excess N<sub>2</sub>O into the atmosphere relative to class A. For the region corresponding to class C (Figure 6; Station LF), more transient, weaker, and shorter-lasting water column stratification is typical. This allows for more effective outgassing during spring and summer compared to the previous two classes. Subsequently, mixing of the water column due to the earlier breakdown of stratification during September-October results in high fluxes of N<sub>2</sub>O into the atmosphere, leaving water either at equilibrium or weakly undersaturated with respect to the atmosphere until the following spring (see Figure 5, left). Finally, class D is typically represented by shallow coastal areas and is characterized by a permanently mixed water column (Figure 6; Station SNS). Here, saturation of N2O increases from spring (specifically, April in case of SNS station), and is maintained at relatively high levels throughout the summer months, eventually decreasing from late summer to early autumn.

Within temporally stratified regions (classes A and B), bulk  $N_2O$  production mostly takes place below the mixed layer, allowing for the gradual buildup of oversaturation (excess  $N_2O$ ) below the pycnocline in waters isolated from the surface mixed layer and without contact with the atmosphere until stratification starts to break down (Figure 7; Stations CCS and NNS). However, in spring some  $N_2O$  production within these regions takes place at the base of mixed layer, fueling the concurrent increase in air-sea fluxes (see Figure 5 for the corresponding stations).

Within class C, when water column is mixed during spring-early summer and early autumn,  $N_2O$  production occurs uniformly throughout the water column (Figure 7; Station LF). Throughout the relatively short stratification period, excess  $N_2O$  builds up below the pycnocline, driving an increase in outflux to the atmosphere in early autumn (see Figure 5). Within permanently mixed regions corresponding to class D, lack of stratification allows for the excess  $N_2O$  to escape to the atmosphere as soon as it is produced (Figure 7; Station SNS). As a result of the SOM analysis, air-sea flux of  $N_2O$  within the Norwegian Trench was attributed to class D, otherwise characteristic for shallow permanently mixed regions. This is explained by a strong density gradient between ~20–30 m depth, leading to the bulk of  $N_2O$  production, followed by outgassing, to take place within the upper mixed layer during the productive season. This results in a temporal pattern of







emissions showing lower rates during winter and higher fluxes in summer (Figure S1 in the supporting information).

#### 4.2. The Northwest European Shelf as a Source of $N_2O$

By synthesizing observational evidence, Bange (2006) estimated mean  $N_2O$  saturations of 113% for the whole European shelf. Our results indicate a mean  $N_2O$  saturation of 104.9% for the northwest European continental shelf (Table 1), which is somewhat lower than Bange (2006) estimate which included the high contributions of the Baltic Sea and some areas of the Mediterranean. For the Central North Sea, Law and Owens (1990) reported saturations of 102.2%, while Bange et al. (1996) gave estimates of 104%, which are rather close to the mean given by our model results for classes B and C, representing the central North Sea region.

With the mean air-sea fluxes of  $2.03 \pm 1.93 \,\mu$ mol N<sub>2</sub>O m<sup>-2</sup> day<sup>-1</sup>, our model results show that the northwest European shelf N<sub>2</sub>O emissions contribute 0.02224 Tg N yr<sup>-1</sup> to the atmosphere. This comprises between 3.3% and 6.8% of the total emissions from European shelves and estuaries which are estimated to be 0.33–0.67 Tg N yr<sup>-1</sup> (Bange, 2006). Almost half (49%) of the total emissions within our study area originate from productive coastal regions corresponding to class D; the highest fluxes occur during summer and autumn,





**Figure 7.** Seasonal distributions of N<sub>2</sub>O production (nmol  $L^{-1} day^{-1}$ ) over a climatological year (based on 2006–2015) at stations CCS (Central Celtic Sea; 49.39°N,8.58°W), NNS (Northern North Sea; 57.25°W,2.5°E), LF (Long Forties area; 57.25°N,0.5°W), and SNS (Southern North Sea; 53°N,2.5°E), representing classes A, B, C, and D, respectively. White contour lines indicate density anomaly (kg m<sup>-3</sup>).

accounting for up to 71% of the total emissions (Table 1). Class D shows the highest emissions (0.00459 Tg N) in summer, regions of C and B classes in autumn (0.00206 and 0.00183 Tg N, respectively), and class A in winter (0.00138 Tg N), following the differences in intensity and duration of stratification among these classes (see Figure 5). Notably, in contrast to other regions and seasons, class C on average acts as a weak

Modeled Daily Mean Air-Sea Fluxes ( $\mu$ mol N<sub>2</sub>O m<sup>-2</sup> day<sup>-1</sup>), Saturations (%) and Total Emissions (Tg N per Season and per Year) From Regions Corresponding to the Identified Classes (A–D)

V	,					
Region/class		А	В	С	D	Total
Mean saturation (%) Mean air-sea flux (umol $m^{-2} d^{-1}$ )		$102.83 \pm 2.25$ 1 57 + 1 79	$102.86 \pm 2.58$ 1 59 + 1 93	$104.56 \pm 4.54$ $1.87 \pm 2.07$	$107.21 \pm 5.79$ 2 59 + 1 93	$104.89 \pm 4.16$ 2 03 + 1 93
		1.57 1.75	1.55 1.55	1.07 <u>-</u> 2.07	2.55 - 1.55	2.05 - 1.55
Total emission, (Ig N)	Spring	0.00052	0.00032	0.0006	0.00258	0.00401
	Summer	0.00077	0.00072	0.00185	0.00459	0.00794
	Autumn	0.00115	0.00183	0.00206	0.00275	0.00779
	Winter	0.00138	0.00024	-0.0001	0.00099	0.0025
	Total $(y^{-1})$	0.00382	0.00311	0.00441	0.01091	0.02224



sink for atmospheric  $N_2O$  in winter (see also Figure 4). Previous modeling work by Holt et al. (2012) has projected that a significant area of the northwest European shelf will become increasingly stratified under the influence of future climate. From the perspective of dynamics of  $N_2O$  fluxes, this is expected to lead to an expansion of class A and B regions in the future, with potential decrease in emissions within these regions due to decreased productivity. This will in turn lead to stronger decoupling from coastal regions which are less influenced by the open ocean processes, i.e. class C and D regions, where mixing is expected to intensify further, along with biogeochemical cycling enhanced by higher temperatures.

As part of its international commitments, the UK submits annual estimates of anthropogenic greenhouse gas emissions to the United Nations Framework Convention on Climate Change (UNFCCC) and the European Union Monitoring Mechanism Regulation (MMR). According to the UK National Atmospheric Emissions Inventory (NAEI, https://naei.beis.gov.uk/), total anthropogenic emissions from all sources within the UK comprise 0.04598 Tg N yr<sup>-1</sup> on average for the period 2006–2015. Our study shows that the on-shelf part of the UK EEZ annually emits 0.00877 Tg N, a considerable addition of 19.1% to the total land-based emission estimates, and 29% compared to UK emissions from agricultural sources, which comprise 0.03027 Tg N yr<sup>-1</sup>. Although N<sub>2</sub>O is naturally produced within the marine environment and therefore is not included in the national accounting, anthropogenic nutrient input has the potential to increase emissions within coastal seas. Further regional modeling efforts are necessary to quantify the contribution of human activity to N<sub>2</sub>O emissions from coastal regions and inform policy on their management potential, for example, via nutrient input reduction measures.

As is the case for the northwest European shelf as a whole, a substantial degree of heterogeneity in spatiotemporal distributions of emissions exists within the on-shelf part of the UK EEZ, as indicated by average air-sea flux of  $1.7 \pm 1.84 \,\mu\text{mol} \,\text{m}^{-2} \,\text{day}^{-1}$  and saturation of  $104.4 \pm 3.86\%$ . This variability highlights that more accurate measurement-based estimates of N<sub>2</sub>O fluxes require higher temporal and spatial sampling resolution, which can be facilitated via the establishment of a harmonized N<sub>2</sub>O Ocean Observation Network, as proposed by Bange et al. (2019). This initiative will support the development of models such as the one used in the current study, for further improvement of our understanding of the present dynamics and future trends in marine N<sub>2</sub>O emissions.

#### 4.3. Critical Assessment of Model Assumptions and the Future Outlook

The modeled  $N_2O$  air-sea flux values (Table 1) are prone to uncertainties in the adopted formulation and parameterization of  $N_2O$  production and air-sea exchange. For instance, the model overestimates  $N_2O$  percent saturation at water depths of ~40 m during summer (Figure 2, right), which potentially results in overestimated sea-to-air flux within temporally stratified regions. Nevertheless, since accumulation of  $N_2O$  in the water column and timing of its release to the atmosphere are controlled by the stratification regime, the reliable representation of seasonality depends on the accuracy of our hydrodynamic model simulations.

Our model formulation allows for the enhancement of  $N_2O$  production at low oxygen conditions. However, within this study the impact on production was negligible due to high oxygen content throughout the model domain (not shown). Our results thus indicate that  $N_2O$  production within the waters of northwest European shelf is primarily a function of substrate (ammonium) availability, and may be modified by temperature and pH variability (Beman et al., 2011; Huesemann et al., 2002; Kitidis et al., 2011) affecting nitrification rate, as parameterized in ERSEM (Butenschön et al., 2016).

As demonstrated by Zamora and Oschlies (2014), nitrification rates and the corresponding  $N_2O$  production within the surface ocean are the major sources of uncertainty for emission estimates within global model settings. Although in our model we did not explicitly limit nitrification (and hence  $N_2O$  production) within the photic zone, the model showed that most of the  $N_2O$  production within temporally stratified regions (classes A and B) occurs at the base of the pycnocline as it starts to develop in spring. During the remainder of the stratified period  $N_2O$  production continued below the pycnocline, with production rates within surface layers remaining very low both during stratified and mixed seasons. These distributions of production correlate with those of ammonium accumulation (Figure S2 in the supporting information). The model therefore suggests that lower nitrification in the photic zone is an emergent property related to the substrate dynamics rather than to photoinhibition. This complements the findings of Smith et al. (2014), who ascribed the competition between nitrifiers and phytoplankton for ammonium as the dominant factor limiting  $N_2O$ 



production within the photic zone. In shallower, well-mixed regions (class D and, to some extent, class C) the supply of ammonium is abundant throughout water column, as is the production of  $N_2O$ , during the entire biologically active period. Nevertheless, the existing uncertainty surrounding the role of light in regulating nitrification rates creates an imperative for further modeling and experimental studies exploring the sensitivity of ammonia oxidation to different light levels and regimes, and its effects on the production of  $N_2O$ .

The model parameterization adopted in this paper focuses on nitrification as a single process generating  $N_2O$ , which, as we have demonstrated, is appropriate for the geographical region and the spatial resolution of the model used in this study. However, its application in areas experiencing lower oxygen concentrations (e.g., upwelling regions) or dynamic sediment activity (e.g., estuaries) requires more detailed consideration of the whole nitrogen cycle. This is particularly so with respect to the parameterization of denitrification within benthic and pelagic zones which may provide either a source or sink of  $N_2O$ . Future iterations should also consider the inclusion of anammox and dissimilatory nitrate reduction to ammonium (DNRA), which may affect the  $N_2O$  cycle by modifying substrate availability. The understanding of these processes, however, remains incomplete, which hinders their implementation within large-scale model simulations. For instance, there is evidence from terrestrial environments that some organisms responsible for DNRA can potentially consume  $N_2O$  (Sanford et al., 2012). The priority for future model development and its implementation within the northwest European shelf is an exploration of the model sensitivity to varied parameterizations of benthic  $N_2O$  sinks and sources, as this has the potential to increase model accuracy in biogeochemically complex productive shallow regions, where the model to data match is weaker (see Figure 3).

Observations of  $N_2O$  concentrations and air-sea exchange remain sparsely distributed both in space and in time, even for the northwest European shelf. Recent developments in technology for increasing the frequency of measurements (e.g., Grefe & Kaiser, 2014) coupled with improvements in flux estimation (e.g., Holding et al., 2019) and systematically coordinated observations (Bange et al., 2019; Wilson et al., 2018) will each contribute to refinements in the  $N_2O$  budget. In addition to the space- and time-resolving measurements of  $N_2O$  concentrations, model developments will require better constraints on the rates and distributions of nitrogen cycling processes over the estuarine-coastal-offshore continuum. The targeting of specific undefined processes, for example, the factors controlling the balance between anammox and denitrification (Kitidis et al., 2017) and the potential for  $N_2O$  consumption in oxygenated waters (e.g., Raes et al., 2016) remain a high priority in order to further our understanding of the extremely complex marine nitrogen cycle and thus to enable the mechanistic implementation of enigmatic processes into future model environment.

# 5. Conclusions

Within this study we have augmented a well-established biogeochemical-ecological model with  $N_2O$ -related processes to study distributions and controls of air-sea fluxes of this potent greenhouse gas within the northwest European continental shelf region.

Our study has shown that physical constraints directly regulate seasonal patterns of  $N_2O$  fluxes on the shelf. The duration and intensity of water column stratification defines the vertical position and timing of  $N_2O$  production. This, in turn, offers the control to the accumulation of excess  $N_2O$  within the water column and the timing and rate of subsequent release to the atmosphere. The physical regime and its inherent seasonal variability also regulate biological productivity which controls the supply of ammonium for nitrification via planktonic excretion and exudation, and via organic matter mineralization, and determines the total amount of  $N_2O$  produced and eventually lost to the atmosphere.

We estimate that the northwest European shelf as a whole is a source of  $N_2O$  to the atmosphere, contributing 0.02224 Tg N annually, which, however, comprises only between 3.3 and 6.8% of total emissions from European shelves and estuaries. Since 49% of the total emissions within our study area originate from coastal regions, changes in anthropogenic nutrient input have the potential to substantially impact on regional  $N_2O$  emissions. Further modeling efforts will quantify the anthropogenic footprint of these emissions and project their future changes.



#### Acknowledgments

This work was supported by project "Detection and Attribution of Regional greenhouse gas Emissions in the UK" (DARE-UK) funded by the UK Natural Environment Research Council (NERC) (Grant NE/S004947/1), project "Radiatively active gases from the North Atlantic Region and Climate Change" (RAGNARoCC) funded by NERC (Grants NE/K002473/1, NE/K002511/1), the UK Shelf Seas Biogeochemistry programme (Grants NE/K001876/1, NE/K002058/1) jointly funded by NERC and the Department for Environment, Food and Rural Affairs (DEFRA), UK Ocean Acidification programme (Grant NE/H017259/1), jointly funded by the DEFRA, NERC, and the Department for Energy and Climate Change (DECC), NERC National Capability grant to the UK Earth System Modelling project (Grant NE/N018036/ 1) and Climate Linked Atlantic Sector Science (CLASS) NERC National Capability Long-term Single Centre Science Programme (Grant NE/R015953/1). L. P. was part funded by the NERC project "Understanding and modeling the Microbial Carbon Pump under changing nutrient concentration and temperature" (Grant NE/R011087/1). This work used the ARCHER UK National Supercomputing Service (http://www. archer.ac.uk). The model ERSEM is freely available at this site (https:// www.pml.ac.uk/Modelling\_at\_PML/ Models/ERSEM). Nitrous oxide measurement data used for the model validation is freely available from the British Oceanographic Data Centre

(https://www.bodc.ac.uk/).

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