

MarLIN Marine Information Network

Information on the species and habitats around the coasts and sea of the British Isles

Novocrania anomala and *Protanthea simplex* on very wave-sheltered circalittoral rock

MarLIN – Marine Life Information Network Marine Evidence-based Sensitivity Assessment (MarESA) Review

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Please note. This MarESA report is a dated version of the online review. Please refer to the website for the most up-to-date version [https://www.marlin.ac.uk/habitats/detail/1162]. All terms and the MarESA methodology are outlined on the website (https://www.marlin.ac.uk)

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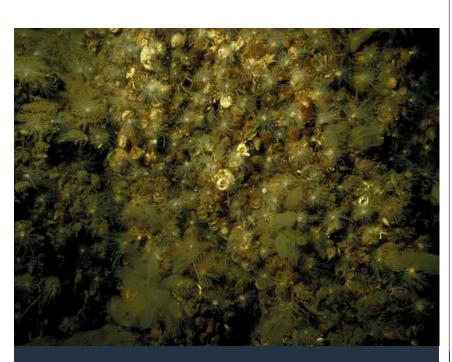
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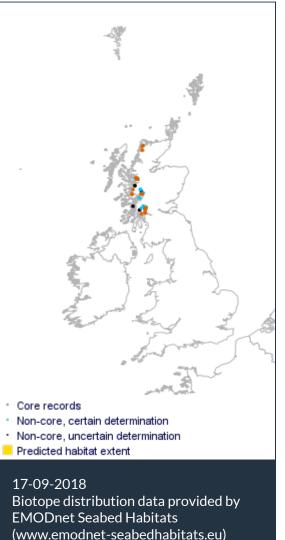
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Neocrania anomala and Protanthea simplex on very wavesheltered circalittoral rock Photographer: Sue Scott Copyright: Joint Nature Conservation Committee (JNCC)



Researched by John Readman Refereed by Admin

Summary

UK and Ireland classification

EUNIS 2008	A4.3141	Neocrania anomala and Protanthea simplex on very wave- sheltered circalittoral rock
JNCC 2015	CR.LCR.BrAs.NovPro.FS	Novocrania anomala and Protanthea simplex on very wave- sheltered circalittoral rock
JNCC 2004	CR.LCR.BrAs.NeoPro.FS	Neocrania anomala and Protanthea simplex on very wave- sheltered circalittoral rock
1007 Distance		

1997 Biotope

Description

This biotope typically occurs on deep, lower circalittoral bedrock or boulder slopes (often-vertical walls) in the landward, very sheltered basins of fjordic sealochs. In these very sheltered conditions, there are frequently dense populations of the anemone *Protanthea simplex* growing on the silty boulder or rock slope, and on the tubes of the parchment worm *Chaetopterus variopedatus*. The

underlying rock surfaces are usually covered with encrusting red algae, the polychaete Spirobranchus triqueter, the brachiopod Novocrania anomala, the saddle oyster Pododesmus patelliformis and the conspicuous fan worm Sabella pavonina. Scattered colonies of Alcyonium digitatum are occasionally present along with the hydroid Bougainvillia ramosa. The barnacle Balanus balanus and the hermit crab Pagurus bernhardus is occasionally seen on boulder or rock surface, whilst underneath in crevices, the squat lobster Munida rugosa may be present. A diverse range of solitary ascidians, typically found in sheltered conditions, are often present including Ciona intestinalis, Corella parallelogramma, Polycarpa pomaria, Ascidia mentula and Ascidia virginea. Echinoderms such as brittlestars Ophiothrix fragilis are frequently seen with their arms protruding from crevices in the rock, whilst the starfish Asterias rubens, the sea urchin Echinus esculentus and Psammechinus miliaris are occasionally found on the boulder or rock surface. The whelk Buccinum undatum is often present but in very low numbers. Where this biotope occurs on vertical rock cliffs, you tend to find silted sugar kelp Saccharina latissima communities above NeoPro, in the infralittoral zone (Slat). Where NeoPro occurs on a rock or boulder slope and due to the very sheltered nature of the habitat, it is common to find a mud plain community where slope ends (CMU). Species such as the seapen Funiculina quadrangularis, the anemone Cerianthus *lloydii* and the Norway lobster *Nephrops norvegicus* are typically abundant. (Information from Connor et al., 2004; JNCC, 2015).

↓ Depth range

5-10 m, 10-20 m, 20-30 m, 30-50 m

<u>m</u> Additional information

-

Listed By

- none -

% Further information sources

Search on:

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Sensitivity review

Sensitivity characteristics of the habitat and relevant characteristic species

CR.LCR.BrAs.NovPro and its sub-biotopes CR.LCR.BrAs.NovPro.FS and CR.LCR.BrAs.NovPro.VS occur on steep or vertical sheltered bedrock and boulder slopes in the circalittoral, typically in fjordic sealochs. CR.LCR.BrAs.NovPro experiences full or variable salinity, and its two sub-biotopes differ in salinity; the low to full salinity NovPro.VS (salinity between <18-35 psu) and the full salinity NovPro.FS (salinity of between 30-35 psu).

The brachiopod Novocrania anomala (previously Neocrania anomala) is the characterizing species for this group. The anemone Protanthea simplex is a characterizing species of the full salinity NovPro.FS, but is only occasionally seen in the variable salinity NovPro.VS. In the variable salinity sub-biotope (NovPro.VS), a diverse range of ascidians, including Dendrodoa grossularia, Ciona intestinalis and Ascidia mentula become more dominant. Sarcodictyon roseum is present in the CR.LCR.BrAs.NovPro.VS, although the description and species list only record species as occasionally present and is considered where appropriate. Other species present in these biotopes are considered transient, mobile or ubiquitous and are, therefore, not considered significant to assessment of the sensitivity of these biotopes. Hence, sensitivity assessments focus on Novocrania anomala, Protanthea simplex and the ascidians (including Dendrodoa grossularia, Ciona intestinalis and Ascidia mentula).

Resilience and recovery rates of habitat

Novocrania anomala (previously Neocrania anomala) is an inarticulate brachiopod that cements its lower shell to the hard substrata (and is therefore sessile) following a pelagic larval stage (Nielsen, 1991, Alvarez & Emig 2000). Brachiopods tend to grow quickly initially, in order to increase survival in the early, most vulnerable life stage. Growth subsequently becomes more stable, diminishing in the latest stages. In general, brachiopods to live up to about 10 years (Alvarez & Emig, 2000). Novocrania anomala possesses no pedicle, cementing its ventral valve directly to the substratum instead and orientating with the dorsal side up, the anterior margin directed upwards away from the substratum (Ruppert & Barnes, 1994). The species is free-spawning and fertilization is external in the surrounding water column. The eggs are denser than seawater and hatch into a free-swimming larval stage. The larvae are fully developed within three days and settle out in no more than a few days, limiting the dispersal range. Although the species may inhabit areas with water flow rates of up to 3 knots, it is often restricted to sheltered habitats such as sea lochs, which may reduce dispersal ability (Jackson, 2000). No information was available about fecundity. Reproduction occurs annually and over an extended period of time (Long & Stricker, 1991; James et al., 1992) with spawning reported between April and October in the South of France and Scotland (Joubin, 1886; Rowell, 1960 cited in James et al., 1992). Novocrania anomala is also capable of recovery from considerable damage to the shell and soft tissue, the adults can be maintained quite well in aquaria and are generally hardy organisms (James et al., 1992).

Protanthea simplex is a small (2 cm high), delicate anemone (Jackson, 2008) commonly found in Scottish lochs (Wood, 2005), which appears to be the southern limit to this species' distribution (NBN, 2016), although one record exists off Connemara, Ireland (Seeley, 2006). This anemone is sociable and beds of up to 2000 per m² are found in Scandinavia (Wood, 2005). *Protanthea simplex* has a long-lived pelagic larval stage (15-20 days, at 10-12°C, in the plankton before settling) and therefore has considerable dispersal potential, with breeding taking place between September and October in Sweden (Jackson, 2008). Fragments of tissue in this species (except the tentacles) are

capable of regenerating into complete anemones, a form of vegetative, asexual reproduction (Manuel, 1988). Anemones are not completely sessile and are capable of slow movement. For example, Sebens (1981) observed immigration to cleared patches of the long-lived anemone *Anthopleura xanthogrammica* as the primary driver towards recovery. Sebens (1981) cleared intertidal patches of *Anthopleura xanthogrammica* at Mukkaw Bay, WA observing the effects over four years. Even after four years, cleared areas were not back to pre-removal population densities. Chia & Spaulding (1972) studying the anemone *Tealia crassicornis* found no sign of gonad development at 14 months old.

Solitary ascidians are discrete creatures which do not fuse with others (unlike colonial ascidians), but may still form dense beds (e.g. up to 5000 individuals/m¹ for *Ciona intestinalis*) (Naylor, 2011). Dendrodoa grossularia, Ascidia mentula and Ciona intestinalis occur across the western, northern and southern coasts of the UK, with more scattered records on the eastern coast (NBN, 2015). Dendrodoa grossularia is a small solitary ascidian (1.5-2 cm diameter (Miller, 1954)). Settlement occurs from April-June, by the following summer individuals reach their maximum size. Life expectancy is expected to be 18-24 months. Sexual maturity is reached within the second year of growth and the release of gametes occurs from spring-autumn, with peaks in early spring and another in late summer. Gamete release is reduced at temperatures above 15°C and totally suppressed above ca. 20°C (Miller, 1954). Kenny & Rees (1994) observed Dendrodoa grossularia was able to recolonize rapidly following aggregate dredging. Following experimental dredging of a site off the English coast, which extracted an area of 1-2 m wide and 0.3-0.5 m deep, Dendrodoa grossularia was able to recolonize and attained 40% of pre-dredge abundance and 23% of biomass within 8 months. This recover rate combined with the ability of this species to reach sexual maturity within its first year suggests that Dendrodoa grossularia can recover from disturbance events within two years.

In Ciona intestinalis, spawning has been reported as year-round in temperate conditions (MBA, 1957, Yamaguchi, 1975, Caputi et al., 2015) with seasonal spawning observed in colder climates from May to June on the Canadian coast (Carver et al., 2006) and in shallower habitats in Sweden (Svane & Havenhand, 1993). Oviparous solitary ascidians generally spawn both oocytes and sperm into the water column, where the resultant fertilized eggs develop into free-swimming, nonfeeding larvae. The eggs are negatively buoyant and slightly adhesive and are either released freely or in mucus strings that are especially adhesive. These strings have a tendency to settle close to or on the parent ascidian. In vitro studies conclude that fertilization proceeds normally whether in the water column or attached to the mucus string. The hatched free-swimming larvae settle nearby, are held by the mucus string until settlement or escape as plankton. Retention in the mucus string may explain the dense aggregations of adults found (Svane & Havenhand, 1993). In vitro studies indicate that both spawning and settlement are controlled by light. However, Ciona intestinalis has been observed in vivo to spawn and settle at any time of the day (Whittington, 1967; Svane & Havenhand, 1993). In the Mediterranean, population collapses of Ciona intestinalis were observed, followed by recovery within 1-2 years (Caputi et al., 2015). The collapses are still poorly understood, although low salinity (Pérès, 1943) and temperature (Sabbadin, 1957) were suggested as possible drivers.

Ascidia mentula is a larger (up to 18 cm long) and longer lived (up to seven years) ascidian compared to *Ciona intestinalis* (Rowley, 2008). Recruitment was reported to occur year round in Sweden at depths greater than 20 m, with seasonal spawning occurring at 15 m (where sea temperature variability is much greater). Both active larvae settlement distribution and passive deposition of larvae (i.e. purely hydrodynamic processes) have been proposed (Havenhand & Svane, 1991 see also Meadows & Campbell, 1972; Scheltema, 1974; Butman, 1987). Long-term data from

populations of the ascidian *Ascidia mentula* on subtidal vertical rock indicated that recruitment of *Ascidia mentula* larvae was positively correlated with adult population density, and then by subsequent active larval choice at smaller scales. Factors influencing larval settlement have been listed as light, substratum inclination and texture (Havenhand & Svane, 1989). Sebens (1985, 1986) described the recolonization of epifauna on vertical rock walls. Rapid colonizers such as encrusting corallines, encrusting bryozoans, amphipods and tubeworms recolonized within 1-4 months. Ascidians such as *Dendrodoa carnea*, *Molgula manhattensis* and *Aplidium* spp. achieved significant cover in less than a year, and, together with *Halichondria panicea*, reached pre-clearance levels of cover after two years. A few individuals of *Alcyonium digitatum* and *Metridium senile* colonized within four years (Sebens, 1986) and would probably take longer to reach pre-clearance levels.

Sarcodictyon roseum forms small colonies on rock surfaces and occasionally on shell, with polyps that reach up to 1 cm in height. It is found from the intertidal to ca 100m and from the Mediterranean to the North Sea (Bay-Nouailhat, 2007; Hayward & Ryland, 1995b). It is found across the western coasts of the British Isles from the Channel Isles to the north coast of Scotland (NBN, 2016) and, more widely, from the Mediterranean to Sweden (Dyntaxa, 2013; van Ofwegen et al., 2001, cited from van Ofwegen, 2015). Very little information is available for this species.

Resilience assessment. The ascidians are likely to recover from any level of mortality relatively quickly, however, *Novocrania anomala* and *Protanthea simplex* are longer lived, slower growing and are unlikely to fully recover from significant mortality within two years. When resistance is None or Low, resilience is classed as '**Medium**', but when resistance is Medium or High, resilience is classed as '**High**'.

🏦 Hydrological Pressures

	Resistance	Resilience	Sensitivity
Temperature increase	Low	<mark>Medium</mark>	<mark>Medium</mark>
(local)	Q: High A: Medium C: Medium	Q: Medium A: Medium C: Medium	Q: Medium A: Medium C: Medium

Novocrania anomala extends as far south as the Mediterranean. Radolović *et al.* (2015) describe the gregarious settlement of *Novocrania anomala* in a cave in the Adriatic coast of Croatia. Despite this geographical range, Hiscock (1985 cited in Hartnoll, 1998), noted that seasonal shallow thermoclines may form, particularly in sheltered areas such as sea lochs, and extend down to 15 m. Some animals such as the brachiopods *Crania* (syn. *Novocrania*) and *Terebratulina* seem restricted to below this thermocline. It is, therefore, possible that geographical populations may acclimate to local temperature (Hartnoll, 1998).

Protanthea simplex is a northerly species, with its southern distribution limit typically in Scotland, although one record exists off Connemara, Ireland (Seeley, 2006). *Protanthea simplex* experienced a marked long-term decline in Swedish and Norwegian sites between 1972 and 1982, experiencing ca 50% mortality (density of ca 600 to ca 300) at 20 m and virtual population collapse at 15 m (reduction in density from 400-500 to <50) over ten years (Heip *et al.*, 1985). This decline appeared to coincide with a warm period between 1973 and 1978 of up to ca +2°C deviation and an increase in *Novocrania anomala* (Heip *et al.*, 1985).

Ciona intestinalis is considered a cold water or temperate species but has been found as far south as Cape Verde, although these tropical populations are likely transitory (Monniot & Monniot, 1994). Temperature tolerance varies among geographical populations or ecotypes. Adult *Ciona intestinalis*

is reported as tolerant of temperatures up to 30°C (Dybern, 1965; Therriault & Herborg, 2008), although Petersen & Riisgard (1992) noted that filtration rates declined above 21°C, which suggested thermal stress, and indicated that long-term survival was likely to require temperatures lower than the 30°C. Other studies also indicated that *Ciona intestinalis* exhibits a decline in ammonia excretion rate and oxygen consumption rate above 18°C (Zhang and Fang 1999, Zhang *et al.*, 1999).

The effect of higher temperatures on *Ascidia mentula* is not as well researched. It is distributed from Norway through to the Mediterranean and the Black Sea, and the species appears to tolerate a broad range of temperatures. Svane (1984) found that in Sweden, whilst lower temperatures decreased recruitment, populations responded positively to the "warm period" of 1972-1976 (Glantz, 2005), with an increase in population density across all sites in the study and a gradual decrease during the ensuing "cold period", and minor fluctuations throughout. Unusually high mean temperatures in 1975 did result in higher recruitment, with colder temperatures in January 1976 and spring 1979 coinciding with very little recruitment. Svane (1984) found that, unlike recruitment, mortality was regulated locally and independent of temperature within the range of the study (mean monthly deviation of $\pm 3^{\circ}$ C)(Svane, 1984).

Sensitivity assessment. Whilst *Novocrania anomala* and the ascidians are unlikely to be affected by a temperature increase at the benchmark level, *Protanthea simplex*, already at its southern-most limit, is unlikely to tolerate a long-term increase in temperature e.g. 2°C for a year. Population collapse has been recorded at 15 m in Sweden during an extended warm period (longer than the benchmark level). Resistance has therefore been assessed as '**Low**', Resilience has been assessed as '**Medium**' and sensitivity is therefore recorded as '**Medium**'.

Temperature decrease	High	High	Not sensitive
(local)	Q: Medium A: Medium C: Medium	Q: High A: High C: High	Q: Medium A: Medium C: Medium

The characterizing brachiopod *Novocrania anomala* has been recorded as far north as Svalbard (Greig, 1924, cited in Prestrud *et al.*, 2004) and the anemone *Protanthea simplex* is a northerly species, occurring in Scandinavia and recorded in across Scotland (Wood, 2005; NBN, 2015). A reduction in temperature would probably be beneficial and could result in distribution expansion of *Protanthea simplex*.

Tolerance for low temperatures varies among geographical populations of ascidians. In the Mediterranean, most adult *Ciona intestinalis* die when temperatures fall below 10°C, and the population is maintained by the survival of younger individuals, which are more tolerant of colder temperatures (Marin *et al.*, 1987). Observation of Scandinavian populations indicated a higher mortality rate of *Ciona intestinalis* during the coldest period of the year (temperatures down to 1°C) (Dybern, 1965).

In Scandinavian populations, normal egg development requires 8-22°C and larval development occurs between 6-24°C (Dybern, 1965). Larval temperature tolerances may play a part in successful recruitment in unseasonable temperature fluctuations. *Ciona savigny* larvae were found to acclimate to temperature, with embryos collected in the summer dividing normally between 14 - 27°C and embryos collected in the winter dividing normally between 10 - 20°C (Nomaguchi *et al.*, 1997).

Ascidia mentula is distributed from Norway through to the Mediterranean and the Black Sea, and the species appears to tolerate a broad range of temperatures. Svane (1984) found that in

Sweden, whilst lower temperatures (of ±3°C of the monthly mean) decreased recruitment, mortality did not significantly increase. Shallow populations (15 m) experiencing much greater seasonal variability did exhibit seasonal spawning rather than year-round spawning that occurs in more temperate and deeper populations (Svane, 1984).

Populations responded positively to the 'warm period' of 1972-1976 (Glantz, 2005), with an increase in population density across all sites in the study and a gradual decrease during the ensuing 'cold period', with minor fluctuations throughout. Unusually high mean temperatures in 1975 did result in higher recruitment, with colder temperatures in January 1976 and spring 1979 coinciding with very little recruitment. Svane (1984) found that, unlike recruitment, mortality was regulated locally and independent of temperature within the range of the study (mean monthly deviation of $\pm 3^{\circ}$ C).

Sensitivity assessment. None of the characterizing species for this biotope are at their northern distribution limit and are unlikely to be affected by a reduction in temperature at the benchmark level. Resistance is therefore assessed as '**High**', Resilience as '**High**' and Sensitivity is therefore recorded as '**Not Sensitive**'.

Salinity increase (local)

No evidence (NEv) Q: NR A: NR C: NR No evidence (NEv) Q: NR A: NR C: NR No evidence (NEv) Q: NR A: NR C: NR

Novocrania anomala occurs in biotopes ranging across all salinity regimes – from Low (<18 ppt) through to Full (30-35 ppt). No literature could be found relating to this brachiopod in hypersaline conditions. *Protanthea simplex* prefers full salinity habitats over lower salinity as, whilst it is characterizing in NovPro.FS, it is only occasionally found in the variable salinity biotope NovPro.VS (Connor *et al.*, 2004). No evidence could be found for the effect of hypersaline conditions.

Ciona intestinalis has been classified as euryhaline with a high salinity tolerance range (12-40‰) although it typically occurs in full salinity conditions (>30‰) (Tillin & Tyler-Walters, 2014). *Ciona intestinalis* has been found in salinities ranging from 11 to 33 PSU in Sweden, although the same study found that parent acclimation to salinity (high or low) has an overriding and significant effect on larval metamorphic success, independent of parent origins (Renborg, 2014).

Sensitivity assessment. NovPro.FS is found in full salinity, but **'No evidence**' for the characterizing species in hypersaline conditions could be found.

Salinity decrease (local)

Medium

High

Low

Q: Medium A: Medium C: Medium Q: Medium A: Medium C: Medium Q: Medium A: Medium C: Medium

Novocrania anomala occurs in biotopes ranging across all salinity variants, from Low (<18 ppt) through to full (30-35 ppt). Whilst Novocrania anomala and the ascidians are likely to be unaffected by a decrease in salinity at the benchmark level, the reduced presence of Protanthea simplex from the variable salinity NovPro.VS indicates that it is at least partially intolerant of salinity decrease. It should also be noted that, given the biotope descriptions, a change from NovPro.FS to NovPro.VS would occur in the event of a permanent salinity reduction at the benchmark level. Whilst no evidence could be found directly for Novocrania anomala, Hammond (1983) reported that Lingula anatina could tolerate salinities ranging from 20‰ to 50‰ for prolonged periods (more than four weeks) and survived in salinities as low as 5‰ for short periods of time (snap response failed at 1 day) and 11.5‰ for 19 days. Other studies report tolerances of ca 16 to 18‰

(Emig, 1997 and references therein) reported more conservative tolerance. It should be noted that the ability for brachiopods to respire anaerobically within their closed shell would enable them to survive short-term changes (James et al., 1992). Ciona intestinalis has been classified as euryhaline with a high salinity tolerance range (12-40‰) although it typically occurs in full salinity conditions (>30‰) (Tillin & Tyler-Walters, 2014) but has been found in Scandinavian waters in salinities as low as 11 PSU (Renborg, 2014, Dybern, 1967). Adult acclimation to salinity was shown to have an overriding and significant effect on larval metamorphic success, independent of parent origins (Renborg, 2014). 'Massive die-offs' of Ciona intestinalis were observed following winter rains in Californian harbours with 'massive recolonizations usually following in the spring' (Lambert & Lambert, 1998). Population collapses of *Ciona intestinalis* in the Mediterranean have also been reported, and whilst the drivers for these events are not well understood, it has been postulated that low salinity could play a part (Péres, 1943; Caputi et al., 2015). Oxygen consumption rate has been shown to decline with decreasing salinity and ceased at 19‰ with siphons tightly closed. (Shumway, 1978). Ascidia mentula is found on the West coast of Norway in salinities greater than 20‰ (Dybern, 1969) and found in a brackish lake in Corsica with a salinity gradient of 6.5 to 18.5‰ Cl⁻ (Verhoeven, 1978). Dendrodoa was observed as one of the dominant species in a study area in Lübeck Bay, Norway, where salinity was recorded as between 11.1 -15.0‰ (Gulliksen, 1977).

Sensitivity assessment. The abundance of *Protanthea simplex* is reduced from NovPro.FS (Full salinity of 35 ppt) to NovPro.VS (Low- Variable salinity, <18– 35ppt) while the abundance of *Dendrodoa grossularia* increases (JNCC, 2014). Therefore, resistance is assessed as '**Medium**, resilience as '**High**' and sensitivity as '**Low'**.

Water flow (tidal current) changes (local)

<mark>High</mark> Q: Medium A: High C: High <mark>High</mark> Q: High A: High C: High

Not sensitive Q: Medium A: High C: High

The CR.LCR.BrAs.NovPro biotope exists in weak to negligible tidal flow conditions (<0.5 m/sec.) Connor *et al.* (2004). High flow rates may reduce the abundance of *Novocrania anomala* and *Protanthea simplex*, although *Novocrania anomala* may inhabit areas with water flow rates of up to 3 knots (Jackson, 2000; 2008b). Decreases in water flow are unlikely to have any effect but increases in flow rate above weak may prevent the animals from maintaining posture and interfere with feeding. Increased flow rates may also sweep individuals off the substratum.

As sessile filter feeders, ascidians generally require a reasonable water flow rate in order to ensure sufficient food availability. It was shown that in stagnant water, phytoplankton density became reduced in a 20-30 cm layer immediately above a dense colony of *Ciona intestinalis* (Riisgård *et al.*, 1996). However, *Ciona intestinalis* has been recognised as tolerant of low water flow environments which it uses as a competitive advantage in areas with minimal water exchange and renewal such as harbours, marinas and docks, (Carver *et al.*, 2006).

Whilst *Ciona intestinalis* is typically found in areas of low flow, it can reportedly withstand flow rates up to 3 knots (1.5 m/s) (Jackson 2008b). If dislodged, juveniles and adults have a limited capability to re-attach, given calm conditions and prolonged contact with the new substrata (Millar 1971; Carver *et al.*, 2006; Jackson, 2008). Hiscock (1983) found that, for the solitary ascidian *Ascidia mentula*, siphons closed when current velocity rose above about 0.15 m/s.

Sensitivity assessment. Significant increases in water flow are likely to negatively affect the characterizing species, but mortality at the benchmark level is unlikely. Therfore, resistance is

assessed as '**High**', resilience as '**High**' and the biotope is considered '**Not sensitive**' at the pressure benchmark.

Emergence regime changes

Not relevant (NR) Q: NR A: NR C: NR Not relevant (NR) Q: NR A: NR C: NR Not relevant (NR) Q: NR A: NR C: NR

Changes in emergence are **not relevant** to this biotope as it is restricted to fully subtidal/circalittoral conditions - the pressure benchmark is relevant only to littoral and shallow sublittoral fringe biotopes.

Wave exposure changes	High
(local)	Q: Low A: NR C: NR

High Q: High A: High C: High Not sensitive

Q: Low A: Low C: Low

The NovPro biotope complex is found in sheltered to extremely sheltered wave exposure conditions. *Novocrania anomala* is unlikely to tolerate wave action and is found in sheltered. Increases in wave exposure above moderately exposed would probably cause death (Jackson, 2000).

High energy wave action can be detrimental to ascidian populations. This is mainly through physical damage to the sea squirts and through the abrasive action of suspended sediment (Jackson, 2008). *Ciona intestinalis* is often dominant in highly sheltered areas such as harbours (Carver *et al.*, 2006). Decreases in wave exposure are unlikely to have any effect. If dislodged, juvenile and adult *Ciona intestinalis* have a limited capability to re-attach, given calm conditions and prolonged contact with the new substratum (Millar 1971; Carver *et al.*, 2006; Jackson 2008;) but increases in wave exposure above moderately exposed are likely to cause a proportion of the population to die, especially in the shallower examples of the biotope if the cobbles and pebbles on which the biotope occurs are mobilized by wave action. *Ascidia mentula* has rarely been recorded at depths shallower than 15 m (Svane, 1984), it is possible that damage could occur if subjected to increased wave exposure.

Sensitivity assessment. Whilst the characterizing species are likely to be affected by an increase in wave exposure, the NovPro biotope complex is found in sheltered to extremely sheltered wave exposure conditions and a change at the benchmark level is unlikely to increase mortality. Resistance is therefore assessed as '**High**', with resilience as '**High**' and the biotope is therefore assessed as '**Not sensitive'** at the benchmark level.

A Chemical Pressures

	Resistance	Resilience	Sensitivity
Transition elements & organo-metal	Not Assessed (NA)	Not assessed (NA)	Not assessed (NA)
contamination	Q: NR A: NR C: NR	Q: NR A: NR C: NR	Q: NR A: NR C: NR

Mercier *et al.* (1998) studied response to TBT exposure of temperate anemones (*Metridium senile* and *Bunodactis stella*), which metabolised and regulated butyl-tin uptake, and accumulated less than mussels (Fent, 1996) or symbiotic anemones (Mercier *et al.*, 1996). Regulation of butyl-tin compounds in temperate sea anemones seemed to parallel that of trace metals (Mercier *et al.*, 1998). Nevertheless, this pressure is **Not assessed** but evidence is presented where available.

Hydrocarbon & PAH contamination

Not Assessed (NA) Q: NR A: NR C: NR Not assessed (NA) Q: NR A: NR C: NR

Not assessed (NA) Q: NR A: NR C: NR

This pressure is **Not assessed** but evidence is presented where available.

CR.LCR.BrAs.NovPro is a sub-tidal biotope (Connor *et al.*, 2004). Oil pollution is mainly a surface phenomenon its impact upon circalittoral turf communities is likely to be limited. However, as in the case of the *Prestige* oil spill off the coast of France, high swell and winds can cause oil pollutants to mix with the seawater and potentially negatively affect sublittoral habitats (Castège *et al.*, 2014). Filter feeders are highly sensitive to oil pollution, particularly those inhabiting the tidal zones that experience high exposure and show correspondingly high mortality, as are bottom dwelling organisms in areas where oil components are deposited by sedimentation (Zahn *et al.*, 1981). Smith (1968) studied the aftermath of the *Torrey Canyon* oil spill and subsequent use of dispersants in that area. The beadlet anemone *Actinia equine* and the dahlia anemone *Tealia feline* were considered some of the most resistant animals on the shore, being commonly found alive, and on 26th April they were found in pools between the tide-marks which appeared to be devoid of all other animals. Some *Anemonia sulcata*, *Sagartia elegans* and *Cereus pedunculatus* were found dead; few survived. Little evidence for brachiopod resistance to hydrocarbon contamination could be found.

Synthetic compound contamination	Not Assessed (NA)	Not assessed (NA)	Not assessed (NA)
	Q: NR A: NR C: NR	Q: NR A: NR C: NR	Q: NR A: NR C: NR
This pressure is Not a	issessed but evidence is p	resented where available.	
Radionuclide	No evidence (NEv)	No evidence (NEv)	No evidence (NEv)
contamination	q: NR A: NR C: NR	q: NR A: NR C: NR	q: NR A: NR C: NR
No evidence			
Introduction of other substances	Not Assessed (NA)	Not assessed (NA)	Not assessed (NA)
	Q: NR A: NR C: NR	Q: NR A: NR C: NR	Q: NR A: NR C: NR
This pressure is Not a	issessed.		
De-oxygenation	<mark>Medium</mark>	<mark>High</mark>	<mark>Low</mark>
	Q: High A: Low C: Medium	Q: Medium A: Medium C: Medium	Q: Medium A: Low C: Medium

The brachiopod *Terebratulina unguicula* is dominant in a fjord that shows annual upwelling of anoxic waters; it is one of the few animals found in areas with oxygen levels frequently below 0.1 ml/l (Tunnicliffe & Wilson, 1988). The inarticulate *Crania californica* has a restricted distribution that may be a function of low dispersal. Over a number of years, one monitored brachiopod population was stable although a gradual retreat from the low oxygen waters was observed (Tunnicliffe & Wilson, 1988).

The brachiopod *Terebratulina septentionalis* has been shown to survive in anoxic conditions for 3.5 days (through anaerobic respiration), however, its activity may be reduced (Hammen, 1977, cited

in James et al., 1992). Thayer (1981) found that the brachiopods *L. californianus* and *Terebratulina unguicula* survived more than two weeks in anoxic conditions.

Ellington(1982) reviewed the effects of hypoxia and anoxia on anemones and noted that survival was several days to weeks depending on species. Mazouni *et al.* (2001) noted that whilst oysters (*Magallana gigas*) can survive short-term exposure to periods of anoxia (Thau Lagoon, France), the associated biofouling community dominated by *Ciona intestinalis* suffered heavy mortality. It should be noted, however, that this species is frequently found in areas with restricted water renewal where oxygen concentrations may drop (Carver *et al.*, 2006). While adverse conditions could affect health, feeding, reproductive capability and could eventually lead to mortality, recovery should be rapid.

In general, respiration in most marine invertebrates does not appear to be significantly affected until extremely low concentrations are reached. For many benthic invertebrates, this concentration is about 2 ml/l (Herreid, 1980; Rosenberg *et al.*, 1991; Diaz & Rosenberg, 1995). Cole *et al.* (1999) suggest possible adverse effects on marine species below 4 mg/l and probable adverse effects below 2 mg/l.

The ability of solitary ascidians to withstand decreasing oxygen levels has not been well documented.

Hiscock & Hoare (1975) reported an oxycline forming in the summer months (Jun-Sep) in a quarry lake (Abereiddy, Pembrokeshire) from close to full oxygen saturation at the surface to <5% saturation below ca 10 m. During these summer events, no echinoderms or *Ascidia mentula* were recorded at depths below 10 - 11 m.

Sensitivity assessment. No direct evidence was found on the effect of hypoxia on *Novocrania anomala* or *Protanthea simplex*. Based on the evidence of other brachiopods, and anemones, a proportion of the community is likely to be able to survive a short-term anoxic episode. An event at the benchmark level is likely to cause limited mortality among all the characterizing species. Resistance is, therefore recorded as '**Medium**', resilience as '**High**' and sensitivity as '**Low**'.

Nutrient enrichment

Not relevant (NR) Q: NR A: NR C: NR Not relevant (NR) Q: NR A: NR C: NR

Not sensitive Q: NR A: NR C: NR

James *et al.* (1992) conducted a review of brachiopod literature, including brachiopod sensitivity to turbidity and cited Rhodes & Thayer (1991), highlighting the ability of brachiopods to sort and reject particles. The rejected material was removed using mucous streamers, although the extra energy cost of this mechanism was noted. The importance of dissolved organic matter is debated (see James et al., 1992). Steele & Petrovic (1975, 1976, 1979) and Thayer (1986) discussed the importance of the unfused lophophore tentacles in adaptation to highly turbid environments, although whether this makes them better suited to excessive turbidity than bivalves is questioned (Thayer, 1991). *Ascidia mentula* has been reported in Iberian bays subject to both nutrient-rich upwelling events and anthropogenic pollution (Aneiros *et al.*, 2015). There is some suggestion that there are possible benefits to ascidians from the increased organic content of water; ascidian 'richness' in Algeciras Bay was found to increase in higher concentrations of suspended organic matter (Naranjo *et al.*, 1996). Nevertheless, this biotope is considered to be '**Not sensitive'** at the pressure benchmark, that assumes compliance with good status as defined by the WFD.

Organic enrichment

No evidence (NEv) Q: NR A: NR C: NR Not relevant (NR) Q: NR A: NR C: NR

No evidence (NEv) Q: NR A: NR C: NR

Rhodes & Thayer (1991, cited in James *et al.*,1992) highlighted the ability of brachiopods to sort and reject particles. The rejected material was removed using mucous streamers, although the extra energy cost of this mechanism was noted. The importance of dissolved organic matter is debated (see James et al., 1992). The lophophore in some brachiopods appears to be able to absorb directly dissolved organic matter from seawater (Storch & Welsch, 1976 cited in Emig, 1997). Steele & Petrovic (1975, 1976, 1979) and Thayer (1986) discussed the importance of the unfused lophophore tentacles in adaptation to highly turbid environments, although whether this makes them better suited to excessive turbidity than bivalves is questioned (Thayer, 1991).

Shick (2012) reviewed the uptake of dissolved organic matter in sea anemones and noted (citing Robbins & Shick, 1980) that this feeding mechanism is likely to particularly important for small anemones, which have a proportionally large tentacle surface area.

There is some suggestion that there are possible benefits to the ascidians from the increased organic content of water. Ascidian 'richness' in Algeciras Bay was found to increase in higher concentrations of suspended organic matter (Naranjo *et al.* 1996). Kocak & Kucuksezgin (2000) noted that *Ciona intestinalis* was one of the rapid breeding opportunistic species that tended to be dominant in Turkish harbours enriched by organic pollutants and was frequently found in polluted environments (Carver *et al.*, 2006). *Ascidia mentula* has been reported in Iberian bays subject to both nutrient-rich upwelling events and anthropogenic organic pollution (Aneiros *et al.*, 2015).

Sensitivity assessment. All of the characterizing species are filter feeders and dissolved organic matter is an important food source. However, organic enrichment can lead to eutrophication or hypoxia. But '**No evidence**' as to the effects of organic enrichment in sublittoral rock was found.

A Physical Pressures



All marine habitats and benthic species are considered to have a resistance of '**None**' to this pressure and to be unable to recover from a permanent loss of habitat (resilience is '**Very low'**). Sensitivity within the direct spatial footprint of this pressure is, therefore '**High**'. Although no specific evidence is described confidence in this assessment is '**High**', due to the incontrovertible nature of this pressure.

Physical change (to another seabed type)

None Q: High A: High C: High Very Low Q: High A: High C: High High Q: High A: High C: High

The replacement of rock by sediment would be a fundamentally change the physical character of the biotope, which would be lost. Resistance to the pressure is considered '**None**', and resilience '**Very low'** (a permanent change). Sensitivity has been assessed as '**High**'.

Novocrania anomala and Protanthea simplex on very wave-sheltered circalittoral rock - Marine Life Information Network

Date: 2018-02-16

Physical change (to another sediment type)

Not relevant (NR) Q: NR A: NR C: NR Not relevant (NR) Q: NR A: NR C: NR

Not relevant (NR) Q: NR A: NR C: NR

'Not relevant' to biotopes occurring on bedrock.

Habitat structure	Not relevant (NR)	Not relevant (NR)	Not relevant (NR)
changes - removal of			
substratum (extraction)	Q: NR A: NR C: NR	Q: NR A: NR C: NR	Q: NR A: NR C: NR

The species characterizing this biotope are epifauna or epiflora occurring on rock and would be sensitive to the removal of the habitat. However, extraction of rock substratum is considered unlikely and this pressure is considered to be '**Not relevant**' to hard substratum habitats.

Abrasion/disturbance of the surface of the	Low	Medium	Medium
substratum or seabed	Q: Low A: NR C: NR	Q: Medium A: Medium C: Medium	Q: Low A: Low C: Low

Novocrania anomala is a sessile brachiopod protected by a calcified shell. However its shell is not massively strong and physical disturbance (such as a passing scallop dredge) will probably cause damage and death (Long & Stricker, 1991; James *et al.*, 1992; Jackson, 2000). *Protanthea simplex* is delicate and soft-bodied. Abrasion is highly likely to cause mortality. Both *Ciona intestinalis* and *Ascidia mentula* are large, emergent, sessile ascidians, and physical disturbance is likely to cause damage and mortality. Emergent epifauna are generally very intolerant of disturbance from fishing gear (Jennings & Kaiser, 1998). However, studies have shown *Ascidia spp.* to become more abundant following disturbance events (Bradshaw *et al.*, 2000).

Sensitivity assessment. Whilst the hard shell of *Novocrania anomala* would afford some protection from a lighter abrasion event, both the ascidians and the soft-bodied *Protanthea simplex* are likely to be significantly affected. Resistance is, therefore, assessed as '**Low**', resilience as '**Medium**', and sensitivity as '**Medium**'.

Penetration or
disturbance of the
substratum subsurfaceNot relevant (NR)Not relevant (NR)Not relevant (NR)Q: NR A: NR C: NRQ: NR A: NR C: NRQ: NR A: NR C: NRQ: NR A: NR C: NR

The species characterizing this biotope group are epifauna or epiflora occurring on rock which is resistant to subsurface penetration. The assessment for abrasion at the surface only is therefore considered to equally represent sensitivity to this pressure. This pressure is thought '**Not Relevant**' to hard rock biotopes

Changes in suspended solids (water clarity)

<mark>High</mark> Q: High A: Medium C: Medium

High Q: High A: High C: High Not sensitive Q: High A: Medium C: Medium

Rhodes and Thayer's (1991, cited in James *et al.*, 1992) highlighted the ability of brachiopods to sort and reject particles. The rejected material was removed using mucous streamers, although the extra energy cost of this mechanism was noted. The importance of dissolved organic matter as a food source was debated (see James *et al.*, 1992). Steele-Petrović (1975, 1976, 1979) and Thayer

(1986) discussed the importance of the unfused lophophore tentacles in adaptation to turbid environments, although whether this makes them better suited to excessive turbidity than bivalves was questioned (Thayer, 1991). Whilst brachiopods appear well suited to cope with turbid conditions, Rudwick (1965, 1970, cited in James et al., 1992) noted that the sessile brachiopods, unable to maintain themselves at the surface of a rapidly accreting substratum, would not tolerate sedimentation. Increased siltation may clog the anemone's tentacles and interfere with feeding. Clearing the sediment will require increased energetic expenditure and loss of condition may result. Ciona intestinalis frequently occurs in habitats close to harbours and marinas with high levels of silt and suspended matter (Carver et al., 2006; Kocak & Kucuksezgin, 2000). Naranjo et al. (1996) described Ciona intestinalis as having a large body and siphons that have wide apertures that helps prevent blocking. Increased suspended sediment may potentially have some detrimental effects in clogging up feeding filtration mechanisms, however, there are possible benefits from increased suspended sediment as ascidian 'richness' in Algeciras Bay was found to increase in higher concentrations of suspended organic matter (Naranjo et al. 1996). In high (up to 300 mg/l of inorganic and 2.5×10^7 cells/l) suspended particulate concentrations, the active rejection mechanism (squirting) is increased in Ciona intestinalis with no discrimination between organic and inorganic particulates observed in any of the ascidians in the study (Robbins, 1984a). Despite these observations, the turbidity tolerance level for this species is not well established. Robbins (1985a) found that continual exposure to elevated levels of inorganic particulates (>25 mg/l) arrested the growth rate of Ciona intestinalis and exposure to 600 mg/l resulted in 50% mortality after 12-15 days and 100% mortality after 3 weeks. It was suggested that because this species can only "squirt" to clear the branchial sac, it may be vulnerable to clogging under heavy sediment loads. Ascidia mentula has been shown to decrease absolute (instantaneous) rate of pumping in high suspended particulate concentrations, whilst filtration efficiency remained unchanged (Robbins, 1984a). However, specific data on the sensitivity to suspended sediment is lacking.

Sensitivity assessment. Whilst increase in turbidity may result in increased energy expenditure for some of the characterizing species, all have methods to cope with increased turbidity. Assuming the biotope typically occurs in intermediate turbidity (10-100 mg/l), an increase at the benchmark level (to medium, 100-300 mg/l) is unlikely to cause significant mortality amongst the characterizing species and Resistance is assessed as '**High**', Resilience as '**High**' and the biotope is considered to be '**Not sensitive**'.

Smothering and siltation Medium rate changes (light) Q: High A: M

Medium Q: High A: Medium C: Medium High

Low

Q: Medium A: Medium C: Medium Q: Medium A: Medium C: Medium

Protanthea simplex is a small (2 cm high), delicate anemone (Jackson, 2008). Smothering with 5 cm of sediment is likely to cause physical damage to the anemone as well as restricting respiration and preventing feeding. In the event of burial, *Novocrania anomala* is up to 1.5 cm in size (Jackson, 2000), can close their shell and survive in low oxygen concentrations, which can be tolerated for a few days (Thayer, 1981 cited in James *et al.*, 1992). Longer term, smothering by sediment will prevent feeding and result in anoxic conditions, eventually resulting in mortality. Adults are sessile, being permanently cemented to the substratum (Jackson, 2000) and Rudwick (1965, 1970, cited in James *et al.*, 1992) noted that the sessile brachiopods, unable to maintain themselves at the surface of a rapidly accreting substratum, would not tolerate sedimentation. The solitary ascidians considered in this report are permanently attached to the substratum and are active suspension feeders. Smothering with 5 cm of sediment is likely to only affect a small proportion of the population because the adults reach up to 15 cm and 18 cm in length for *Ciona intestinalis* and

Ascidia mentula respectively (Rowley, 2008; Jackson, 2008) and frequently inhabit vertical surfaces (Jackson, 2008), Recovery should be rapid, facilitated by the remaining adults.

Sensitivity assessment. Smothering by 5 cm of sediment is likely to impact the characterizing species of this biotope including the sessile ascidians and Novocrania anomala. The biotope often occurs on vertical walls which would protect the characterizing species from sedimentation effects. In flatter areas where sedimentation could occur, given the low water flow rates experienced by these biotopes (Weak to very weak), it is unlikely that the sediment would be removed very rapidly and it is unlikely that Novocrania anomala could survive. Resistance has been assessed as 'Medium', resilience as 'High' and sensitivity has been assessed as 'Low' at the benchmark level.

Smothering and siltation Low rate changes (heavy)

Q: High A: Medium C: Medium

Medium

Medium

Q: Medium A: Medium C: Medium Q: Medium A: Medium C: Medium

Protanthea simplex is a small (2 cm high), delicate anemone (Jackson, 2008). Smothering with 30 cm of sediment is highly likely to cause physical damage to the anemone as well as restricting respiration and preventing feeding. In the event of burial, the dorsal valve of Novocrania anomala is up to 1.5 cm in size (Jackson, 2000), can close their shell and survive in low oxygen concentrations, which can be tolerated for a few days (Thayer, 1981 cited in James et al., 1992). Longer term, smothering by sediment will prevent feeding and result in anoxic conditions, eventually resulting in mortality. Adults are sessile, being permanently cemented to the substratum (Jackson, 2000) and Rudwick (1965, 1970, cited in James et al., 1992) noted that the sessile brachiopods, unable to maintain themselves at the surface of a rapidly accreting substratum, would not tolerate sedimentation. The solitary ascidians considered in this report are permanently attached to the substratum and are active suspension feeders. Smothering with 30 cm of sediment is likely to affect a proportion of the population because, although the adults reach up to 15 cm and 18 cm in length for Ciona intestinalis and Ascidia mentula respectively (Rowley, 2008; Jackson, 2008), they frequently inhabit vertical surfaces (Jackson, 2008). Recovery should be rapid, facilitated by the remaining adults.

Sensitivity assessment. Smothering by 30 cm of sediment is likely to impact the characterizing species of this biotope especially the sessile ascidians and Novocrania anomala. The biotope often occurs on vertical walls which would protect the characterizing species from sedimentation effects. In flatter areas where sedimentation could occur, given the low water flow rates experienced by these biotopes (weak to very weak), it is unlikely that the sediment would be removed very rapidly and few Novocrania anomala and ascidians would survive. The resilience assessment assumes removal of the sediment, however, depending on the conditions, this may not be possible. Resistance has been assessed as 'Low', resilience as 'Medium' and sensitivity has been assessed as 'Medium' at the benchmark level.

L	itte	r

Not Assessed (NA) Q: NR A: NR C: NR

Not assessed (NA) Q: NR A: NR C: NR

Not assessed (NA) Q: NR A: NR C: NR

Not assessed.

Electromagnetic changes No evidence (NEv)

No evidence (NEv) Q: NR A: NR C: NR

No evidence (NEv) Q: NR A: NR C: NR

No evidence

Underwater noise changes

Not relevant (NR) Q: NR A: NR C: NR Not relevant (NR) Q: NR A: NR C: NR

Not relevant (NR) Q: NR A: NR C: NR

No evidence could be found for the effects of noise or vibrations on *Novocrania anomala* or *Protanthea simplex* but they are considered unlikely to be sensitive to this pressure. McDonald (2014) studied the effect of generator noise on fouling of four vessels by *Ciona intestinalis* and found that fouling was highest at locations closest to the generators and lowest furthest away from the generators. Subsequent *in vitro* experiments demonstrated that larvae settled much faster in the presence of noise (2 h- 20 h compared with 6 h-26 h for control), underwent metamorphosis more rapidly (between 10 and 20 h compared with ca 22 h) and had a markedly increased survival rate to maturity (90-100% compared with 66%). Other studies also reported that noise emissions from vessels promoted fouling by organisms including ascidians (Stanley *et al.*, 2016).

Sensitivity assessment: Resistance to this pressure is assessed as '**High**' and resilience as '**High**'. This biotope is therefore considered to be '**Not sensitive**'.

Introduction of light or	<mark>High</mark>	High	Not sensitive
shading	Q: Medium A: Medium C: Medium	Q: High A: High C: High	Q: Low A: NR C: NR

Whilst most brachiopods are believed to detect light, the means of detection has been questioned and few studies have been undertaken (James et al., 1992). It should be noted that McCammon (1973) reported that *Magellania venosa* was not sensitive to light. Anemones have been reported to respond to light (Parker, 1919 cited in Sebens, 1981). The ascidian *Dendrodoa* spp. larvae do not possess an eye spot, and light does not appear to affect locomotion (Mackie & Bone, 1976).

Sensitivity assessment. Whilst there is evidence that some characterizing species respond to light, either as larvae or as adults, change at the benchmark level is unlikely to cause mortality among any of the species assessed, and resistance and resilience are both recorded as '**High**', and the biotope is considered '**Not sensitive**' to this pressure.

Barrier to species movement

Not relevant (NR) Q: NR A: NR C: NR Not relevant (NR) Q: NR A: NR C: NR Not relevant (NR) Q: NR A: NR C: NR

Not relevant, barriers and changes in tidal excursion are not relevant to biotopes restricted to open waters.

Death or injury by collision

Not relevant (NR) Q: NR A: NR C: NR Not relevant (NR) Q: NR A: NR C: NR Not relevant (NR) Q: NR A: NR C: NR

Not relevant to seabed habitats. NB. Collision by grounding vessels is addressed under 'surface abrasion'.

Visual disturbance

Not relevant (NR) Q: <u>NR</u> A: <u>NR</u> C: <u>NR</u> Not relevant (NR) Q: NR A: NR C: NR Not relevant (NR) Q: NR A: NR C: NR

Not relevant

Biological Pressures

	Resistance	Resilience	Sensitivity
Genetic modification & translocation of	No evidence (NEv)	No evidence (NEv)	No evidence (NEv)
indigenous species	Q: NR A: NR C: NR	Q: NR A: NR C: NR	Q: NR A: NR C: NR

Ciona intestinalis is considered a fouling species and adheres readily to the hulls of vessels, and is considered an invasive species in the USA, Chile, Western Australia, New Zealand, Canada and South Africa (Millar 1966; McDonald 2004; Blum *et al.* 2007; Ramsay *et al.* 2008, 2009; Dumont *et al.*,2011). Whilst there have been novel proposals to farm *Ciona intestinalis* as a potential feedstock for aquaculture in Sweden (Laupsa, 2015), there is no evidence to suggest such farming exists. No evidence for *Novocrania anomala* or *Protanthea simplex* could be found. Therefore, there is currently **'No evidence'** on which to assess this pressure.

Introduction or spread or invasive non-indigenous	f No evidence (NEv)	No evidence (NEv)	No evidence (NEv)
species	Q: NR A: NR C: NR	Q: NR A: NR C: NR	Q: NR A: NR C: NR

This biotope is classified as circalittoral and therefore no algal species have been considered. *Didemnum vexillum* is an invasive colonial sea squirt native to Asia which was first recorded in the UK in Darthaven Marina, Dartmouth in 2005. *Didemnum vexillum* can form extensive mats over the substrata it colonizes, binding boulders, cobbles and altering the host habitat (Griffith *et al.*, 2009). *Didemnum vexillum* can also grow over and smother the resident biological community. Recent surveys within Holyhead Marina, North Wales have found *Didemnum vexillum* growing on and smothering native tunicate communities, including *Ciona intestinalis* (Griffith *et al.*, 2009). Due to the rapid-re-colonization of *Didemnum vexillum* eradication attempts have to date failed. Presently *Didemnum vexillum* is isolated to several sheltered locations in the UK (NBN, 2015), however, *Didemnum vexillum* has successfully colonized offshore in Georges Bank, USA (Lengyel *et al.*, 2009). *Styela clava*, another INIS ascidian, was first recorded in the UK at Plymouth in 1952 (Eno *et al.*, 1997).

There is '**No evidence'** at present that this biotope has been affected by INIS, however, *Didemnum vexillum* could pose a potential threat. Due to the constant risk of new invasive species, the literature for this pressure should be revisited.

Introduction of microbia	No evidence (NEv)	No evidence (NEv)	No evidence (NEv)
pathogens	Q: NR A: NR C: NR	Q: NR A: NR C: NR	Q: NR A: NR C: NR

Little is known about brachiopod disease, therefore there is 'No evidence' on which to base an assessment. However, James *et al.* (1992) highlighted significant shell and soft tissue regenerative capabilities. McCammon (1973) recorded infection of the lophophore tentacles of *Magellania venosa* by an unidentified protist which destroyed connective tissues and ultimately resulted in death. McCammon (1972) also observed brachiopods subject to stress being prone to fungal infection which exhibited black spots forming on the shell. More generally, James *et al.* (1992) discussed brachiopod success in aquaria and suggests that they are relatively hardy. No evidence

for diseases affecting *Protanthea simplex* could be found. There appears to be little research into ascidian diseases, particularly in the Atlantic. The parasite *Lankesteria ascidiae* targets the digestive tubes and can cause 'long faeces syndrome' in *Ciona intestinalis* (although it has also been recorded in other species). Mortality occurs in severely affected individuals within about a week following first symptoms. (Mita *et al.*, 2012). There is '**No evidence'** to support an assessment of this pressure.

Removal of targetLowMediumMediumspeciesQ: Low A: NR C: NRQ: Medium A: Medium C: MediumQ: Low A: Low C: Low

Among brachiopods, only the lingulids (*Lingula* spp.) have been fished commercially, and on a very small scale (Printrakoon & Kamlung-ek, 2013). No evidence for brachiopod extraction in the British Isles was found. Whilst unlikely, should targeted extraction occur, it is likely to have a significant impact on the characterizing species. Whilst anemones are not sessile, they are very slow moving and are therefore likely to be as exposed to extraction as the sessile *Novocrania anomala* and ascidians. **'Not relevant'** as none of the characterizing species are targeted.

Removal of non-target species

Low

Q: Low A: NR C: NR

Medium

Medium

Q: Medium A: Medium C: Medium Q: Low A: Low C: Low

The characteristic species probably compete for space within the biotope, so that loss of one species would probably have little if any effect on the other members of the community. Removal of the characteristic epifauna due to by-catch is likely to remove a proportion of the biotope and change the biological character of the biotope. These direct, physical impacts are assessed through the abrasion and penetration of the seabed pressures. The sensitivity assessment for this pressure considers any biological/ecological effects resulting from the removal of non-target species on this biotope. Therefore, resistance is assessed as 'Low', resilience as 'Medium' and sensitivity assessed as 'Medium'. This biotope may be removed or damaged by static or mobile gears that are targeting other species (see physical pressures above).

Bibliography

Álvarez, F. & Emig, C., 2000. Brachiopoda from the Luso-Iberian zone. I. Biology and ecology. In *The Millennium Brachiopod Congress, London, 2000.* Abstracts.

Aneiros, F., Rubal, M., Troncoso, J.S. & Bañón, R., 2015. Subtidal benthic megafauna in a productive and highly urbanised semienclosed bay (Ría de Vigo, NW Iberian Peninsula). *Continental Shelf Research*, **110**, 16-24.

Bay-Nouailhat, W., 2007. Description de *Sarcodictyon roseum*. [cited 17.03.2016]. Available from: http://www.mer-littoral.org/05/sarcodictyon-roseum.php

Blum, J.C., Chang, A.L., Liljesthröm, M., Schenk, M.E., Steinberg, M.K. & Ruiz, G.M., 2007. The non-native solitary ascidian *Ciona intestinalis* (L.) depresses species richness. *Journal of Experimental Marine Biology and Ecology*, **342** (1), 5-14.

Bradshaw, C., Veale, L.O., Hill, A.S. & Brand, A.R., 2000. The effects of scallop dredging on gravelly seabed communities. In: *Effects of fishing on non-target species and habitats* (ed. M.J. Kaiser & de S.J. Groot), pp. 83-104. Oxford: Blackwell Science.

Butman, C.A., 1987. Larval settlement of soft-sediment invertebrates: the spatial scales of pattern explained by active habitat selection and the emerging role of hydrodynamical processes. *Oceanography and Marine Biology: an Annual Review*, **25**, 113-165.

Caputi, L., Crocetta, F., Toscano, F., Sordino, P. & Cirino, P., 2015. Long-term demographic and reproductive trends in *Ciona intestinalis* sp. A. *Marine Ecology*, **36** (1), 118-128.

Carlgren, O., 1921. Actiniaria. Pt. 1. Danish Ingolf Expedition, Vol. V, No. 9., pp. 31. Copenhagen: Bianco Luno.

Carlgren, O., 1949. A survey of the Ptychodactiaria, Corallimorpharia and Actiniaria. *Kungliga Svenska Vetenskapsakadamiens* Handlingar, Series 4, **1**, 16-110.

Carver, C., Mallet, A. & Vercaemer, B., 2006. Biological synopsis of the solitary tunicate *Ciona intestinalis*. Canadian Manuscript Report of Fisheries and Aquatic Science, No. 2746, v + 55 p. Bedford Institute of Oceanography, Dartmouth, Nova Scotia.

Castège, I., Milon, E. & Pautrizel, F., 2014. Response of benthic macrofauna to an oil pollution: Lessons from the "Prestige" oil spill on the rocky shore of Guéthary (south of the Bay of Biscay, France). *Deep Sea Research Part II: Topical Studies in Oceanography*, **106**, 192-197.

Chia, F.S. & Spaulding, J.G., 1972. Development and juvenile growth of the sea anemone *Tealia crassicornis*. *Biological Bulletin*, *Marine Biological Laboratory*, *Woods Hole*, **142**, 206-218.

Cole, S., Codling, I.D., Parr, W., Zabel, T., 1999. Guidelines for managing water quality impacts within UK European marine sites [On-line]. *UK Marine SACs Project*. [Cited 26/01/16]. Available from: http://www.ukmarinesac.org.uk/pdfs/water_quality.pdf

Connor, D.W., Allen, J.H., Golding, N., Howell, K.L., Lieberknecht, L.M., Northen, K.O. & Reker, J.B., 2004. The Marine Habitat Classification for Britain and Ireland. Version 04.05. ISBN 1 861 07561 8. In JNCC (2015), *The Marine Habitat Classification for Britain and Ireland Version* 15.03. [2019-07-24]. Joint Nature Conservation Committee, Peterborough. Available from https://mhc.jncc.gov.uk/

Connor, D.W., Dalkin, M.J., Hill, T.O., Holt, R.H.F. & Sanderson, W.G., 1997a. Marine biotope classification for Britain and Ireland. Vol. 2. Sublittoral biotopes. *Joint Nature Conservation Committee*, Peterborough, JNCC Report no. 230, Version 97.06., *Joint Nature Conservation Committee*, Peterborough, JNCC Report no. 230, Version 97.06.

Davies, C.E. & Moss, D., 1998. European Union Nature Information System (EUNIS) Habitat Classification. *Report to European Topic Centre on Nature Conservation from the Institute of Terrestrial Ecology, Monks Wood, Cambridgeshire*. [Final draft with further revisions to marine habitats.], Brussels: European Environment Agency.

Diaz, R.J. & Rosenberg, R., 1995. Marine benthic hypoxia: a review of its ecological effects and the behavioural responses of benthic macrofauna. *Oceanography and Marine Biology: an Annual Review*, **33**, 245-303.

Dumont, C., Gaymer, C. & Thiel, M., 2011. Predation contributes to invasion resistance of benthic communities against the non-indigenous tunicate *Ciona intestinalis*. *Biological Invasions*, **13** (9), 2023-2034.

Dybern, B.I., 1965. The life cycle of Ciona intestinalis (L.) f. typica in relation to the environmental temperature. Oikos, 16, 109-131.

Dybern, B.I., 1967. The distribution and salinity tolerance of *Ciona intestinalis* (L.) *f. typica* with special reference to the waters around southern Scandinavia. *Ophelia*, **4** (2), 207-226.

Dybern, B.I., 1969. Distribution and ecology of ascidians in Kviturdvikpollen and Vågsböpollen on the west coast of Norway. *Sarsia*, **37** (1), 21-40.

Ellington, W.R., 1982. Metabolic responses of the sea anemone *Bunodosoma cavernata* (Bosc) to declining oxygen tensions and anoxia. Physiological zoology, 240-249.

Emig, C.C., 1997. Ecology of inarticulated brachiopods. In *Treatise of Invertebrate Paleontology* (Kaesler, RL; editor). Part H, 473-495. Geological Society of America and the University of Kansas Press.

Eno, N.C., Clark, R.A. & Sanderson, W.G. (ed.) 1997. Non-native marine species in British waters: a review and directory. Peterborough: Joint Nature Conservation Committee.

Fent, K., 1996. Ecotoxicology of organotin compounds. *Critical reviews in toxicology*, **26** (1), 3-117.

Glantz, M.H., 2005. Climate variability, climate change and fisheries. Cambridge: Cambridge University Press.

Griffith, K., Mowat, S., Holt, R.H., Ramsay, K., Bishop, J.D., Lambert, G. & Jenkins, S.R., 2009. First records in Great Britain of the

invasive colonial ascidian Didemnum vexillum Kott, 2002. Aquatic Invasions, 4 (4), 581-590.

Gulliksen, B., 1977. Studies from the U.W.L. "Helgoland" on the macrobenthic fauna of rocks and boulders in Lübeck Bay (western Baltic Sea). *Helgoländer wissenschaftliche Meeresunters*, **30**, 519-526.

Hammond, L., 1983. Experimental studies of salinity tolerance, burrowing behavior and pedicle regeneration in *Lingula anatina* (Brachiopoda, Inarticulata). *Journal of Paleontology*, 1311-1316.

Hartnoll, R.G., 1998. Circalittoral faunal turf biotopes: an overview of dynamics and sensitivity characteristics for conservation management of marine SACs, Volume VIII. *Scottish Association of Marine Sciences, Oban, Scotland*. [UK Marine SAC Project. Natura 2000 reports.]

Havenhand, J. & Svane, I., 1989. Larval behaviour, recruitment, and the role of adult attraction in *Ascidia mentula* O. F. Mueller: *Reproduction, genetics and distributions of marine organisms*. 23rd European Marine Biology Symposium. Olsen and Olsen, 127-132.

Havenhand, J.N. & Svane, I., 1991. Roles of hydrodynamics and larval behaviour in determining spatial aggregation in the tunicate *Ciona intestinalis. Marine Ecology Progress Series*, **68**, 271-276.

Hayward, P.J. & Ryland, J.S. (ed.) 1995a. The marine fauna of the British Isles and north-west Europe. Volume 2. Molluscs to Chordates. Oxford Science Publications. Oxford: Clarendon Press.

Heip, C.H., Keegan, B.F. & Lewis, J.R., 1985. Long-Term Changes in Coastal Benthic Communities. In Proceedings of a Symposium, held in Brussels, Belgium, December 9–12, 1985: Springer Science & Business Media.

Herreid, C.F., 1980. Hypoxia in invertebrates. Comparative Biochemistry and Physiology Part A: Physiology, 67 (3), 311-320.

Hiscock, K. & Hoare, R., 1975. The ecology of sublittoral communities at Abereiddy Quarry, Pembrokeshire. *Journal of the Marine Biological Association of the United Kingdom*, **55** (4), 833-864.

Hiscock, K., 1983. Water movement. In Sublittoral ecology. The ecology of shallow sublittoral benthos (ed. R. Earll & D.G. Erwin), pp. 58-96. Oxford: Clarendon Press.

Hiscock, K., 1985. Littoral and sublittoral monitoring in the Isles of Scilly. September 22nd to 29th, 1984. *Nature Conservancy Council, Peterborough*, CSD Report, no. 562., Field Studies Council Oil Pollution Research Unit, Pembroke.

Holt, T.J., Jones, D.R., Hawkins, S.J. & Hartnoll, R.G., 1995. The sensitivity of marine communities to man induced change - a scoping report. *Countryside Council for Wales, Bangor, Contract Science Report*, no. 65.

Jackson, A. 2000. Novocrania anomala, A brachiopod. In Tyler-Walters H. and Hiscock K. (eds) Marine Life Information Network: Biology and Sensitivity Key Information Reviews, [on-line]. Plymouth: Marine Biological Association of the United Kingdom. [17.03.16] Available from: http://www.marlin.ac.uk/species/detail/1331

Jackson, A. 2008b. *Protanthea simplex*, Sealoch anemone. In Tyler-Walters H. and Hiscock K. (eds) *Marine Life Information Network: Biology and Sensitivity Key Information Reviews*, [on-line]. Plymouth: Marine Biological Association of the United Kingdom. [17.03.2016] Available from: http://www.marlin.ac.uk/species/detail/1345

Jackson, A., 2008. Ciona intestinalis. A sea squirt. Marine Life Information Network: Biology and Sensitivity Key Information Subprogramme [On-line]. Plymouth: Marine Biological Association of the United Kingdom. [cited 16/12/15] Available from: http://www.marlin.ac.uk/species/detail/1369

James, M.A., Ansell, A.D., Collins, M.J., Curry, G.B., Peck, L.S. & Rhoda, M.C., 1992. Biology of living brachiopods. Advances in Marine Biology, 28, 175-387.

Jennings, S. & Kaiser, M.J., 1998. The effects of fishing on marine ecosystems. Advances in Marine Biology, **34**, 201-352.

JNCC, 2015. The Marine Habitat Classification for Britain and Ireland Version 15.03. (20/05/2015). Available from https://mhc.jncc.gov.uk/

JNCC, 2015. The Marine Habitat Classification for Britain and Ireland Version 15.03. (20/05/2015). Available from https://mhc.jncc.gov.uk/

JNCC (Joint Nature Conservation Committee), 1999. Marine Environment Resource Mapping And Information Database (MERMAID): Marine Nature Conservation Review Survey Database. [on-line] http://www.jncc.gov.uk/mermaid

Kerby, J.L., Richards-Hrdlicka, K.L., Storfer, A. & Skelly, D.K., 2010. An examination of amphibian sensitivity to environmental contaminants: are amphibians poor canaries? *Ecology Letters*, **13** (1), 60-67.

Kocak, F. & Kucuksezgin, F., 2000. Sessile fouling organisms and environmental parameters in the marinas of the Turkish Aegean coast. *Indian journal of marine sciences*, **29** (2), 149-157.

Lambert, C.C. & Lambert, G., 1998. Non-indigenous ascidians in southern California harbors and marinas. *Marine Biology*, **130** (4), 675-688.

Laupsa, M., 2015. Spawning, settlement and growth of Ciona intestinalis in Øygarden, Hardangerfjorden and Kvitsøy. Master's thesis. University of Bergen.

Lengyel, N.L., Collie, J.S. & Valentine, P.C., 2009. The invasive colonial ascidian *Didemnum vexillum* on Georges Bank - Ecological effects and genetic identification. *Aquatic Invasions*, **4**(1), 143-152.

Long, J.A. & Stricker, S.A., 1991. Brachiopoda. In *Reproduction of marine invertebrates*, Vol. VI. *Echinoderms and Lophophorates*. (ed. A.C. Giese, J.S. Pearse & V.B. Pearse). California: The Boxwood Press.

Mackie, G. & Bone, Q., 1976. Skin impulses and locomotion in an ascidian tadpole. Journal of the Marine Biological Association of the United Kingdom, 56 (03), 751-768.

Magorrian, B.H. & Service, M., 1998. Analysis of underwater visual data to identify the impact of physical disturbance on horse mussel (*Modiolus modiolus*) beds. *Marine Pollution Bulletin*, **36**, 354-359.

Manuel, R.L., 1988. British Anthozoa. London: Academic Press. [Synopses of the British Fauna, no. 18.]

Marin, M.G., Bresan, M., Beghi, L. & Brunetti, R., 1987. Thermo-haline tolerance of *Ciona intestinalis* (L. 1767) at different developmental stages. *Cahiers de Biologie Marine*, **28**, 45-57.

Mazouni, N., Gaertner, J. & Deslous-Paoli, J.-M., 2001. Composition of biofouling communities on suspended oyster cultures: an *in situ* study of their interactions with the water column. *Marine Ecology Progress Series*, **214**, 93-102.

MBA (Marine Biological Association), 1957. *Plymouth Marine Fauna*. Plymouth: Marine Biological Association of the United Kingdom.

McCammon, H.M., 1972. Establishing and Maintaining Articulate Brachiopods in Aquaria. *Journal of Geological Education*, **20** (3), 139-142.

McCammon, H.M., 1973. The ecology of Magellania venosa, an articulate brachiopod. Journal of Paleontology, 266-278.

McDonald, J., 2004. The invasive pest species *Ciona intestinalis* (Linnaeus, 1767) reported in a harbour in southern Western Australia. *Marine Pollution Bulletin*, **49** (9), 868-870.

Meadows, P.S. & Campbell, J.I., 1972. Habitat selection by aquatic invertebrates. Advances in Marine Biology, 10, 271-382.

Mercier, A., Pelletier, É. & Hamel, J.-F., 1998. Response of temperate sea anemones to butyltin contamination. *Canadian Journal of Fisheries and Aquatic Sciences*, **55** (1), 239-245.

Mercier, A., Pelletier, E. & Hamel, J.-F., 1996. Toxicological response of the symbiotic sea anemone *Aiptasia pallida* to butyltin contamination. *Marine Ecology Progress Series*, **14** (1), 133-146.

Millar, R., 1971. The biology of ascidians. Advances in marine biology, 9, 1-100.

Millar, R.H., 1966. Tunicata Ascidiacea. Oslo, Universitetsforlaget.

Mita, K., Kawai, N., Rueckert, S. & Sasakura, Y., 2012. Large-scale infection of the ascidian *Ciona intestinalis* by the gregarine *Lankesteria ascidiae* in an inland culture system. *Diseases of aquatic organisms*, **101** (3), 185-195.

Monniot, C. & Monniot, F., 1994. Additions to the inventory of eastern tropical Atlantic ascidians; arrival of cosmopolitan species. *Bulletin of Marine Science*, **54** (1), 71-93.

Naranjo, S.A., Carballo, J.L., & Garcia-Gomez, J.C., 1996. Effects of environmental stress on ascidian populations in Algeciras Bay (southern Spain). Possible marine bioindicators? *Marine Ecology Progress Series*, **144** (1), 119-131.

Naylor. P., 2011. Great British Marine Animals, 3rd Edition. Plymouth. Sound Diving Publications.

NBN, 2015. National Biodiversity Network 2015(20/05/2015). https://data.nbn.org.uk/

Nielsen, C., 1991. The development of the brachiopod *Crania (Neocrania) anomala* (OF Müller) and its phylogenetic significance. *Acta Zoologica*, **72** (1), 7-28.

Nomaguchi, T.A., Nishijima, C., Minowa, S., Hashimoto, M., Haraguchi, C., Amemiya, S. & Fujisawa, H., 1997. Embryonic thermosensitivity of the ascidian, *Ciona savignyi. Zoological Science*, **14** (3), 511-515.

Nyholm, K-G., 1959. On the development of the primitive actinian *Protanthea simplex*, Carlgren. *Zoologiska Bidrag Fran Uppsala*, Band 33 1958-1962, 69-78.

Pérès, J.M., 1943. Recherches sur le sang et les organes neuraux des Tuniciers. Annales de l'Institut Oceanographique (Monaco), **21**, 229-359.

Petersen, J. & Riisgård, H.U., 1992. Filtration capacity of the ascidian *Ciona intestinalis* and its grazing impact in a shallow fjord. *Marine Ecology-Progress Series*, **88**, 9-17.

Picton, B.E. & Morrow, C.C., 2015. Ascidia mentula O F Müller, 1776. In Encyclopedia of Marine Life of Britain and Ireland. [cited 26/01/16]. Available from: http://www.habitas.org.uk/marinelife/species.asp?item=ZD1500

Prestrud, P., Strøm, H. & Goldman, H.V., 2004. A catalogue of the terrestrial and marine animals of Svalbard. Norsk Polarinstitutt.

Printrakoon, C. & Kamlung-ek, A., 2013. Socioeconomic study and economic value of living fossil, *Lingula* sp. in mangrove ecosystem in Trat Province, Thailand. *Chinese Journal of Population Resources and Environment*, **11** (3), 187-199.

Radolović, M., Bakran-Petricioli, T., Petricioli, D., Surić, M. & Perica, D., 2015. Biological response to geochemical and hydrological processes in a shallow submarine cave. *Mediterranean Marine Science*, **16** (2), 305-324.

Ramsay, A., Davidson, J., Bourque, D. & Stryhn, H., 2009. Recruitment patterns and population development of the invasive ascidian *Ciona intestinalis* in Prince Edward Island, Canada. *Aquatic Invasions*, **4** (1), 169-176.

Ramsay, A., Davidson, J., Landry, T. & Stryhn, H., 2008. The effect of mussel seed density on tunicate settlement and growth for the cultured mussel, *Mytilus edulis*. *Aquaculture*, **275** (1), 194-200.

Renborg, E., Johannesson, K. & Havenhand, J., 2014. Variable salinity tolerance in ascidian larvae is primarily a plastic response to the parental environment. *Evolutionary ecology*, **28** (3), 561-572

Riisgård, H.U., Jürgensen, C. & Clausen, T., 1996. Filter-feeding ascidians (*Ciona intestinalis*) in a shallow cove: implications of hydrodynamics for grazing impact. *Journal of Sea Research*, **35** (4), 293-300.

Robbins, I., 1984a. The regulation of ingestion rate, at high suspended particulate concentrations, by some phleobranchiate ascidians. *Journal of Experimental Marine Biology and Ecology*, **82** (1), 1-10.

Rosenberg, R., Hellman, B. & Johansson, B., 1991. Hypoxic tolerance of marine benthic fauna. *Marine Ecology Progress Series*, **79**, 127-131.

Rowley, S.J., 2008. A sea squirt (*Ascidia mentula*). Tyler-Walters, H. and Hiscock, K. (eds). *Marine Life Information Network: Biology and Sensitivity Key Information Reviews* [on-line]. Plymouth: Marine Biological Association of the United Kingdom. [cited 26/01/16]. Available from: http://www.marlin.ac.uk/species/detail/8

Rudwick, M.J.S., 1970. Living and fossil brachiopods. London: Hutchinson University Library

Ruppert, E.E. & Barnes, R.D., 1994. Invertebrate zoology (6th ed.). Fort Worth, USA: Saunders College Publishing.

Sabbadin, A., 1957. Il ciclo biologico di Ciona intestinalis (L.), Molgula manhattensis (De Kay) e Styela plicata (Lesueur) nella Laguna Veneta.

Scheltema, R.S., 1974. Biological interactions determining larval settlement of marine invertebrates. *Thalassia Jugoslavica*, **10**, 263-296.

Sebens, K.P., 1981. Recruitment in a Sea Anemone Population: Juvenile Substrate Becomes Adult Prey. *Science*, **213** (4509), 785-787.

Sebens, K.P., 1985. Community ecology of vertical rock walls in the Gulf of Maine: small-scale processes and alternative community states. In *The Ecology of Rocky Coasts: essays presented to J.R. Lewis*, D.Sc. (ed. P.G. Moore & R. Seed), pp. 346-371. London: Hodder & Stoughton Ltd.

Sebens, K.P., 1986. Spatial relationships among encrusting marine organisms in the New England subtidal zone. *Ecological Monographs*, **56**, 73-96.

Seeley, R., 2006. Sealife survey. MarLIN news, 9.

Service, M. & Magorrian, B.H., 1997. The extent and temporal variation of disturbance to epibenthic communities in Strangford Lough, Northern Ireland. *Journal of the Marine Biological Association of the United Kingdom*, **77**, 1151-1164.

Service, M., 1998. Recovery of benthic communities in Strangford Lough following changes in fishing practice. *ICES Council Meeting Paper*, CM 1998/V.6, 13pp., Copenhagen: International Council for the Exploration of the Sea (ICES).

Shick, J.M., 2012. A functional biology of sea anemones: Springer Science & Business Media.

Shumway, S., 1978. Respiration, pumping activity and heart rate in *Ciona intestinalis* exposed to fluctuating salinities. *Marine Biology*, **48** (3), 235-242.

Shumway, S.E., 1978. Activity and respiration of the sea anemone, *Metridium senile* (L.) exposed to salinity fluctuations. *Journal of Experimental Marine Biology and Ecology*, **33**, 85-92.

Smith, J.E. (ed.), 1968. 'Torrey Canyon'. Pollution and marine life. Cambridge: Cambridge University Press.

Stanley, J.A., Wilkens, S., McDonald, J.I. & Jeffs, A.G., 2016. Vessel noise promotes hull fouling. In *The Effects of Noise on Aquatic Life II*: Springer, pp. 1097-1104.

Steele-Petrović, H.M., 1975. An explanation for the tolerance of brachiopods and relative intolerance of filter-feeding bivalves for soft muddy bottoms. *Journal of Paleontology*, 552-556.

Steele-Petrović, H.M., 1976. Brachiopod food and feeding processes. Palaeontology, 19 (3), 417-436

Steele-Petrović, H.M., 1979. The physiological differences between articulate brachiopods and filter-feeding bivalves as a factor in the evolution of marine level-bottom communities. *Palaeontology*, **22** (1), 101-134.

Svane, I., 1984. Observations on the long-term population dynamics of the perennial ascidian, *Ascidia mentula* O F Müller, on the Swedish west coast. *The Biological Bulletin*, **167** (3), 630-646.

Svane, I. & Havenhand, J.N., 1993. Spawning and dispersal in *Ciona intestinalis* (L.) *Marine Ecology*, *Pubblicazioni della Stazione Zoologica di Napoli*. I, **14**, 53-66.

Thayer, C.W., 1986. Are brachiopods better than bivalves? Mechanisms of turbidity tolerance and their interaction with feeding in articulates. *Paleobiology*, **12** (02), 161-174.

Therriault, T.W. & Herborg, L.-M., 2008. Predicting the potential distribution of the vase tunicate *Ciona intestinalis* in Canadian waters: informing a risk assessment. *ICES Journal of Marine Science: Journal du Conseil*, **65** (5), 788-794.

Tillin, H. & Tyler-Walters, H., 2014. Assessing the sensitivity of subtidal sedimentary habitats to pressures associated with marine activities. Phase 2 Report – Literature review and sensitivity assessments for ecological groups for circalittoral and offshore Level 5 biotopes. *JNCC Report* No. 512B, 260 pp. Available from: www.marlin.ac.uk/publications

Tunnicliffe, V. & Wilson, K., 1988. Brachiopod populations: Distribution in fjords of British Columbia(Canada) and tolerance of low oxygen concentrations. *Marine ecology progress series*. **47** (2), 117-128.

Van Ofwegen, L., 2015. *Sarcodictyon roseum*. *World Register of Marine Species*: [17/03/16]. Available from http://marinespecies.org/aphia.php?p=taxdetails&id=125340

Van Ofwegen, L., Grasshoff, M. & Van der Land, J., 2001. Octocorallia (excl. Pennatulacea). European register of marine species: a check-list of the marine species in Europe and a bibliography of guides to their identification. Collection Patrimoines Naturels, 50, 104-105.

Veale, L.O., Hill, A.S., Hawkins, S.J. & Brand, A.R., 2000. Effects of long term physical disturbance by scallop fishing on subtidal epifaunal assemblages and habitats. *Marine Biology*, **137**, 325-337.

Verhoeven, J. & Van Vierssen, W., 1978a. Structure of macrophyte dominated communities in two brackish lagoons on the island

of Corsica, France. Aquatic Botany, 5, 77-86.

Whittingham, D.G., 1967. Light-induction of shedding of gametes in *Ciona intestinalis* and *Morgula manhattensis*. *Biological Bulletin*, *Marine Biological Laboratory*, Woods Hole, **132**, 292-298.

Wood. C., 2005. Seasearch guide to sea anemones and corals of Britain and Ireland. Ross-on-Wye: Marine Conservation Society.

WoRMS, 2015. World Register of Marine Species. (11/04/2007). http://www.marinespecies.org

Yamaguchi, M., 1975. Growth and reproductive cycles of the marine fouling ascidians *Ciona intestinalis*, *Styela plicata*, *Botrylloides violaceus*, and *Leptoclinum mitsukurii* at Aburatsubo-Moroiso Inlet (Central Japan). *Marine Biology*, **29** (3), 253-259.

Zahn, R., Zahn, G., Müller, W., Kurelec, B., Rijavec, M., Batel, R. & Given, R., 1981. Assessing consequences of marine pollution by hydrocarbons using sponges as model organisms. *Science of The Total Environment*, **20** (2), 147-169.

Zhang, J. & Fang, J., 1999. Study on the oxygen consumption rates of some common species of ascidian. *Journal of fishery sciences of China*, **7** (1), 16-19.

Zhang, J., Fang, J. & Dong, S., 1999. Study on the ammonia excretion rates of four species ascidian. *Marine Fisheries Research*, **21** (1), 31-36.